

Helena Valley Ground Water: Pharmaceuticals, Personal Care Products, Endocrine Disruptors (PPCPs), and Microbial Indicators of Fecal Contamination



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March, 2006

Montana Bureau of Mines and Geology Open-File Report 532

Abstract

The city of Helena, Montana and its surrounding valley (fig. 1) are experiencing marked population growth with attendant proliferation of onsite wastewater disposal (septic tanks and drainfields) systems. Thirty-eight public and private domestic water supplies deriving ground water from the Quaternary/Tertiary valley-fill aquifer and various bedrock formations were sampled in the summer and fall of 2005 for pharmaceutically active compounds, personal care products, and endocrine disrupting compounds (PPCP as used here).

The two most frequently detected PPCPs are sulfamethoxazole (SMX) and the herbicide atrazine, with detection frequencies of 80% and 40%, respectively. Atrazine demonstrates a strong correlation with chloride and total dissolved solids (TDS). Because chloride and TDS are commonly used inorganic indicators of water-quality degradation from domestic wastewater discharge, the correlation suggests that atrazine could be occurring in domestic wastewater. This hypothesis should be verified in subsequent investigations.

The wells were also sampled for microbial indicators of fecal contamination and for inorganic constituents. There is a poor correlation between the microbial indicators of fecal contamination and PPCP occurrence, with zero detections of either *Escherichia coli* or the somatic or male-specific coliphage. Total coliform was detected at only eight sites.

Introduction

Twenty-two PPCPs have been detected in ground water used for drinking water for private and public water supplies in the Helena valley, Montana. PPCPs are a group of compounds that include antibiotics, hormones, and drugs. Results of several recent studies (Godfrey, 2004; Hinkle, 2005; Heberer, 2004) show that PPCPs are present in relatively low concentrations [nanograms per liter (ng/L) to micrograms per liter ($\mu\text{g/L}$ ranges)] in municipal and domestic wastewater as well as in some surface and ground water. The presence of these compounds in ground water and surface water has drawn public attention not only because of potential health risks from exposure to one or a mixture of these chemicals, but also because the primary mode of entry into our environment is not from manufacturing discharge but from widespread and continual use in human and veterinary and clinical practice (Lancet, 2002) and discharge associated with domestic wastewater. Low levels of various PPCPs in ground water provide clear evidence that domestic wastewater is a source of contamination. In spite of a growing body of evidence describing their distribution in the environment, little is known about their mobility and persistence in ground water or surface water, nor are their effects on human health and aquatic ecosystems well understood.

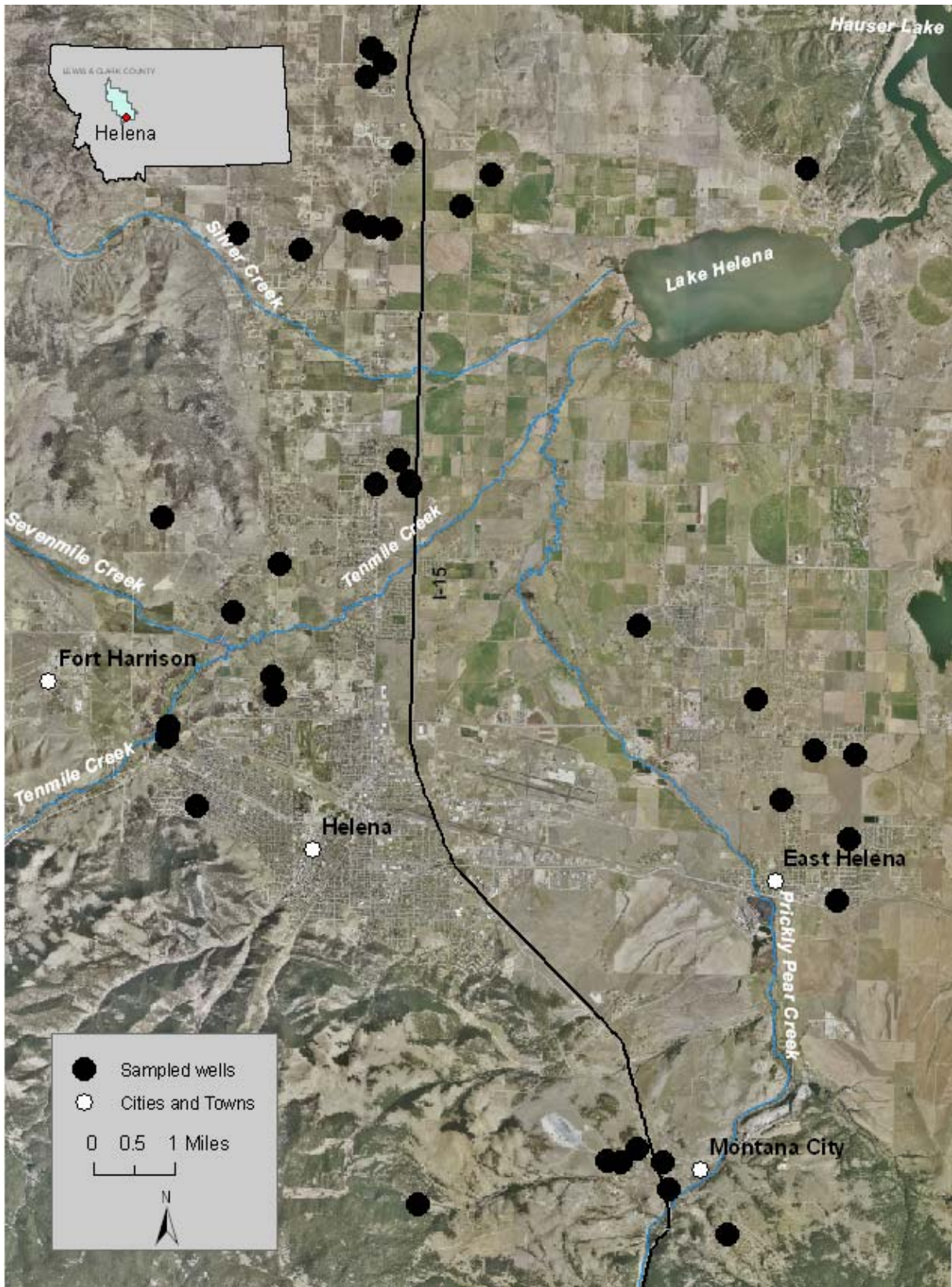


Figure 1. General location of the Helena valley, Montana.

The proposed Ground Water Rule of the National Primary Drinking Water Regulations (40 CFR Parts 141 and 142, May 10, 2000) recognizes the need for a targeted risk-based regulatory strategy that identifies those systems with source-water contamination and systems deriving ground water from hydrogeologically “sensitive” aquifers. Among other stipulations of the proposed Ground Water Rule, public water supplies may be required to monitor ground-water sources for multiple indicators of fecal contamination; under the proposed rule both a bacterium (*Escherichia coli* or enterococci) and a virus (male-specific and somatic coliphage) could be used as indicators. Previous investigators have found coliphage, PPCPs, and other organic wastewater compounds in ground water and in septic tanks (onsite wastewater). In each study the types of analytes differ somewhat. In a shallow unconfined sandy aquifer near La Pine, Oregon, the U.S. Geological Survey (Hinkle, 2005) found 45 organic wastewater compounds in onsite wastewater. In ground-water samples only 9 of the 45 wastewater compounds were found, along with sulfamethoxazole (SMX), acetaminophen, and caffeine. They found that the reactivity of this particular suite of organic wastewater compounds may limit their usefulness as tracers of onsite wastewater discharged into aquifers. In the same study coliphage was frequently detected in onsite wastewater but was only occasionally detected (8 occurrences) at low concentrations in wells, with a consistent absence in replicate or repeat samples. The authors speculate that coliphage was probably attenuated to less than 1 plaque-forming unit (PFU)/100 mL before reaching the sampled wells.

Heberer (2004) noted that more than 60 pharmaceutical residues have been detected in surface water but only a very limited number of the compounds have been found in ground water, and suggests that not only is there a small number of ground-water studies, but also the compounds are likely removed or attenuated during transport into ground water. USGS investigators (2005) found that nitrate and chloride concentrations in onsite wastewater exhibited small variability among systems but that concentrations of individual organic wastewater compounds varied dramatically among different onsite wastewater treatment systems—not uncommonly by several orders of magnitude—suggesting that loading rates of wastewater compounds might be highly variable.

After the analysis of 42 septic tanks and influent to and effluent from the public wastewater treatment facility (WWTF) for Missoula, Montana, Godfrey and Woessner (2004) found 18 pharmaceutically active compounds in septic tanks, 12 in the WWTF influent and 9 in the WWTF effluent. The most frequently detected non-prescription drugs were acetaminophen, caffeine, and nicotine; frequently found prescription drugs were codeine, trimethoprim, and carbamazepine. In a similar evaluation of organic wastewater compounds in septic tanks at about 20 sites in New Jersey, Szabo (2004) found 4-nonylphenol, phenol, caffeine, cotinine, menthol, 3- β -coprostanol, cholesterol, and β -sitosterol.

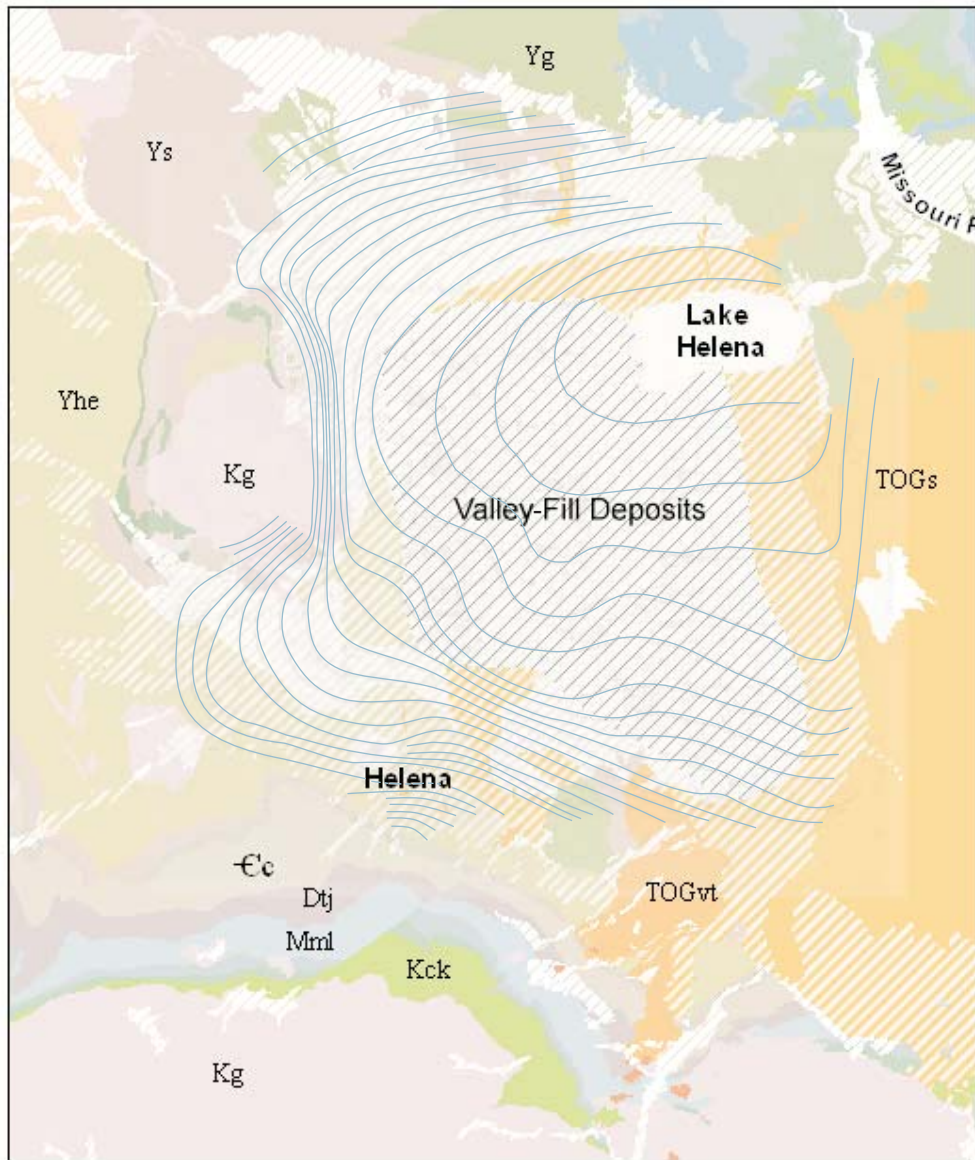
Background

The Helena valley in west-central Montana comprises about 330 square miles (207,400 acres) and is underlain by about 6,000 ft of valley fill composed of Tertiary sediments unconformably overlain by about 100 ft of Quaternary alluvium. Because of the hydraulic interconnection of water-yielding zones, the valley-fill deposits function as one complex aquifer system (Briar, 1992). Surface water enters the valley from Prickly Pear, Tenmile, Sevenmile, and Silver Creeks and from the Missouri River after it has been diverted into irrigation canals. Ground water and surface water discharge principally to Lake Helena and ultimately to the Missouri River. The Helena valley is bounded by folded and fractured sedimentary, metamorphic, and igneous bedrock of Precambrian to Cretaceous age (fig. 2). Figure 2 also shows ground-water-level contours that depict flow from the south, west, and north margins of the valley toward Lake Helena.

Ground-water quality in the valley-fill deposits is characterized as a calcium bicarbonate type with a median pH of 7.5. As shown in Table 1, arsenic, uranium, and nitrate are elevated in ground-water samples at a few locations in the Helena Valley. The maximum values are 17.1 µg/L for arsenic [Maximum Contaminant Level (MCL) = 10 µg/L], 29.1 µg/L for uranium (MCL = 30 µg/L) and 12.4 mg/L for nitrate (MCL = 10 mg/L). Irrigation with arsenic-laden Missouri River water is a possible explanation for the elevated arsenic concentrations. Uraniferous rocks in surrounding bedrock are the probable source of elevated uranium. Elevated nitrate could be indicative of water-quality degradation from domestic wastewater; agricultural sources are possible but less numerous than those derived from domestic wastewater.

Historically a mining and agricultural area, the city of Helena and its surrounding valley are now experiencing dramatic increases in the density of individual onsite wastewater disposal facilities and residential wells. As shown in figure 3, the number of wells installed per year from 1973 through 1994 has averaged about 190. But in the period from 1995–2005 the average number of wells installed escalated to 284 per year. It can be assumed that most of these wells are being installed to serve residences that are not served by city water or sewer services.

A microbial occurrence survey of the Helena valley was conducted in April 2004 by Steve Kilbreath and Joe Meek of the Montana Department of Environmental Quality (DEQ) and Kathy Moore of the Lewis and Clark County Water Quality Protection District (LCWQPD). Results of that survey showed positive male-specific coliphage occurrence in 10 of 19 sampled wells in the Helena valley with no detections of either of *E. coli* or enterococci. Subsequent re-sampling in August 2004 produced negative results for all coliphage, *E. coli*, and enterococci (Steve Kilbreath, Kathy Moore, and Joe Meek, unpub. data, 2004).



Legend

- | | |
|---|---|
| €c Upper and Middle Cambrian carbonate rocks | No bedrock |
| €c | TOGs Oligocene and sedimentary rocks |
| €c | TOGs |
| Dtj Three Forks Formation and Jefferson Formation undivided | TOGvt Oligocene volcanic-stratified tuft |
| Dtj | TOGvt |
| Kck Upper and Lower Cretaceous and sedimentary rocks | Yg Greyson Formation |
| Kck | Yg |
| Kg Cretaceous intrusive rocks, mostly granitic | Yhe Helena and Empire Formation undivided |
| Kg | Yhe |
| Mml Madison Group | Ys Spokane Formation |
| Mml | Ys |
| | Ground-water Level Contours |

Figure 2. Bedrock geology and water-level contours for the Helena valley, Montana. Modified from Thamke and Reynolds (2002) and Briar and Madison (1992).

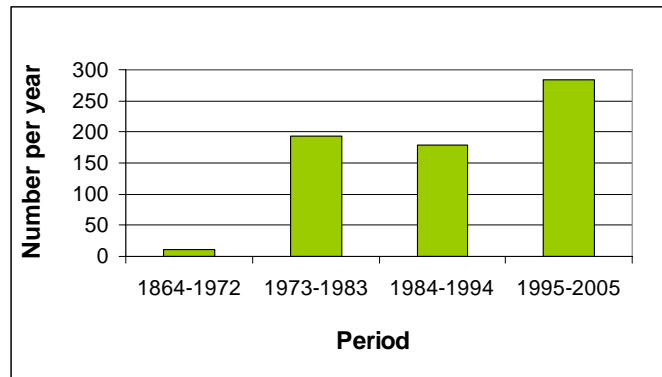


Figure 3. Annual number of new well installations in the Helena valley from 1864 to 2005.

Methods

Thirty-eight wells representing both bedrock ($n=12$) and valley-fill ($n=26$) aquifers were sampled for total coliform, *E. coli*, enterococci, male-specific, and somatic coliphage in April, June, and November 2005. During the same period, 35 wells were sampled for 28 PPCPs and inorganic constituents. Eighteen wells serve small public water supplies with the remainder serving private residences. Well depths range from 39 to 425 ft (table 1).

Wells were flushed prior to sampling until the field parameters of pH, specific conductance, and temperature were stable as per the Montana Bureau of Mines and Geology (MBMG) Standard Operating Procedure for Collection of Ground-Water Samples for Inorganic Analyses (unpub., 2004).

Samples for the analysis of male-specific and somatic coliphage were collected and analyzed in accordance with proposed EPA Method 1601: Male Specific (F+) and Somatic Coliphage in Water by Two-Step Enrichment Procedure (USEPA, 2000). Total coliform, *E. coli*, and enterococci samples were collected and analyzed using Autoanalysis Colilert (MDPHHS, 2004) and Enterolert systems (MDPHHS, 2004), respectively.

Samples for the analysis of PPCPs were collected as grab samples in 1-L amber bottles. After arrival at the lab, Columbia Analytical Services in Kelso, WA prepared the samples using EPA Method 3535 and analyzed the samples using LC/MS/MS (Columbia Analytical Services, 2005).

Samples to be analyzed for inorganic constituents were field-filtered and preserved prior to shipment to the MBMG Analytical Laboratory. The sampling procedures followed the MBMG Standard Operating Procedures for Collection of Ground-Water Samples for Inorganic Analyses (unpub., 2004). The inorganic analytical methods used follow EPA protocols appropriate for the analyte being measured.

Each well site is assigned a unique identification number that can be cross-referenced to the Montana Ground-Water Information Center (GWIC ID). All pertinent well construction, site inventory, and water-quality data may be found on the website, <http://mbmggwic.mtech.edu>.

Table 1. Results of dissolved inorganic analyses for 35 well sites in the Helena Valley with maximum, minimum and median values.

| Gwic Id | Sample Date | Total Well Depth (ft) | Water Temp | Field pH | Field S.C. | Lab pH | Lab Specific Conductance (umhos/cm2) | Total Dissolved Solids (mg/l) | Ca (mg/l) | Mg (mg/l) | Na (mg/l) | K (mg/l) | Fe (mg/l) | Mn (mg/l) | SiO2 (mg/l) | HCO3 (mg/l) | CO3 (mg/l) | SO4 (mg/l) | Cl (mg/l) | NO3 (mg/l) | F (mg/l) | OPO4 (mg/l) |
|---------|----------------|-----------------------|------------|----------|------------|--------|--------------------------------------|-------------------------------|-----------|-----------|-----------|----------|-----------|-----------|-------------|-------------|------------|------------|-----------|------------|----------|-------------|
| 62523 | 5/24/05 | 50 | 10.3 | 7.3 | 376 | 7.85 | 510 | 315.2 | 70.7 | 20.6 | 10.9 | 2.3 | 0.0 | <0.001 | 20.8 | 298.0 | 0.0 | 38.2 | 4.8 | 0.0 | 0.1 | <0.05 |
| 64826 | 5/23/05 | 42 | 10.2 | 7.9 | 285 | 8.19 | 532 | 333.8 | 54.8 | 25.3 | 25.6 | 0.8 | 0.0 | <0.001 | 26.4 | 264.0 | 0.0 | 53.4 | 15.1 | 1.9 | 0.4 | <0.05 |
| 5756 | 5/23/05 | 66 | 10.1 | 7.8 | 625 | 7.92 | 840 | 790.2 | 83.7 | 34.9 | 56.9 | 3.5 | 0.0 | 0.0 | 274.0 | 386.0 | 0.0 | 111.0 | 29.5 | 5.4 | 0.4 | 0.7 |
| 62570 | 5/23/05 | 70 | 12.6 | 7.3 | 1667 | 7.57 | 2580 | 1810.8 | 221.0 | 119.0 | 235.0 | 6.9 | 0.0 | <0.001 | 35.1 | 616.0 | 0.0 | 538.0 | 342.0 | 9.9 | <1.0 | <1.0 |
| 64806 | 5/23/05 | 41 | | 7.8 | 611 | 7.42 | 846 | 528.7 | 47.4 | 21.6 | 115.0 | 1.1 | 0.0 | <0.001 | 38.4 | 421.0 | 0.0 | 71.8 | 17.1 | 8.6 | 0.3 | <0.05 |
| 194850 | 5/24/05 | 180 | 1 | 7.3 | 523 | 7.78 | 702 | 446.3 | 93.4 | 30.8 | 16.2 | 3.3 | 0.0 | <0.001 | 19.6 | 353.0 | 0.0 | 90.8 | 17.6 | 0.7 | <0.05 | <0.05 |
| 62369 | 5/31/05 | 110 | 10.2 | 7.7 | 341 | 7.58 | 838 | 546.5 | 53.1 | 23.5 | 119.0 | 1.2 | 0.0 | 0.0 | 40.8 | 418.5 | 0.0 | 75.4 | 17.6 | 9.5 | 0.2 | <0.05 |
| 62575 | 5/31/05 | 93 | 10.1 | 7.6 | 229 | 7.59 | 863 | 543.3 | 51.9 | 22.8 | 118.0 | 1.2 | 0.0 | 0.0 | 39.9 | 419.9 | 0.0 | 75.4 | 17.6 | 9.4 | 0.2 | <0.05 |
| 65388 | 6/5/05 | 87 | 10.1 | 7.5 | 578 | 7.76 | 587 | 364.0 | 63.8 | 26.0 | 28.9 | 1.7 | 0.0 | <0.001 | 19.3 | 201.1 | 0.0 | 76.1 | 45.2 | 3.9 | 0.0 | 0.0 |
| 170202 | 6/5/05 | 300 | 10.3 | 7.6 | 378 | 7.52 | 508 | 312.5 | 58.2 | 14.5 | 27.9 | 5.6 | 0.0 | 0.0 | 13.7 | 228.1 | 0.0 | 62.7 | 17.1 | 0.0 | 0.4 | <0.05 |
| 187850 | 5/30/05 | 100 | 10.2 | 7.7 | 452 | 7.75 | 607 | 364.5 | 61.8 | 25.2 | 27.2 | 1.7 | 0.0 | <0.001 | 19.0 | 199.3 | 0.0 | 83.9 | 43.5 | 3.8 | 0.2 | <0.10 |
| 206394 | 5/30/05 | 200 | 10.1 | 7.8 | 943 | 7.68 | 1288 | 731.5 | 128.0 | 60.3 | 35.6 | 2.3 | <0.005 | 0.0 | 16.2 | 160.1 | 0.0 | 159.0 | 240.0 | 11.2 | <0.63 | 0.0 |
| 165085 | 7/15/05 | 201 | 10.8 | 7.3 | 252 | 7.77 | 273 | 199.0 | 29.7 | 6.5 | 14.1 | 3.3 | 0.0 | 0.0 | 45.7 | 130.8 | 0.0 | 29.2 | 4.9 | 0.9 | 0.3 | <0.05 |
| 220274 | 7/14/05 | | 12.4 | 7.3 | 617 | 7.64 | 607 | 397.4 | 80.3 | 24.3 | 19.3 | 3.5 | <0.005 | <0.001 | 27.0 | 255.9 | 0.0 | 87.4 | 27.1 | 2.0 | 0.4 | <0.05 |
| 220272 | 7/15/05 | | 12.7 | 7.6 | 689 | 7.94 | 729 | 425.7 | 67.2 | 34.8 | 26.7 | 1.8 | 0.0 | <0.001 | 17.7 | 188.2 | 0.0 | 83.6 | 96.4 | 4.7 | 0.0 | 0.0 |
| 58685 | 7/15/05 | 310 | 17.4 | 7.5 | 504 | 7.78 | 462 | 268.8 | 44.5 | 22.6 | 13.9 | 2.9 | <0.005 | <0.001 | 19.8 | 215.0 | 0.0 | 51.5 | 7.1 | 0.0 | 0.5 | 0.1 |
| 58712 | 7/14/05 | 148 | 12.2 | 7.2 | 902 | 7.41 | 838 | 511.4 | 96.8 | 44.3 | 29.8 | 4.8 | <0.005 | <0.001 | 21.8 | 309.9 | 0.0 | 115.0 | 35.8 | 10.4 | 0.0 | 0.0 |
| 165017 | 7/19/05 | 94 | 12.3 | 7.4 | 444 | 7.85 | 462 | 288.9 | 51.9 | 12.4 | 34.6 | 2.1 | 0.0 | <0.001 | 9.1 | 237.2 | 0.0 | 44.7 | 15.7 | 1.5 | 0.1 | <0.05 |
| 65071 | 7/19/05 | 39 | 10.9 | 7.1 | 470 | 7.54 | 467 | 280.7 | 54.8 | 14.3 | 20.7 | 3.2 | <0.005 | <0.001 | 20.8 | 197.4 | 0.0 | 49.4 | 16.6 | 3.5 | 0.2 | <0.05 |
| 61051 | 7/20/05 | 123 | | | | 7.31 | 451 | 288.8 | 59.1 | 14.7 | 17.3 | 3.2 | 0.0 | <0.001 | 23.1 | 195.4 | 0.0 | 59.6 | 13.4 | 2.0 | 0.2 | <0.05 |
| 61055 | 7/19/05 | 145 | 11.2 | 7.0 | 351 | 7.34 | 395 | 261.9 | 51.6 | 13.2 | 14.8 | 3.1 | 0.0 | <0.001 | 23.5 | 183.7 | 0.0 | 54.4 | 9.4 | 1.2 | 0.2 | <0.05 |
| 62802 | 7/19/05 | 130 | 15.5 | 7.4 | 561 | 7.78 | 543 | 325.7 | 55.3 | 29.5 | 17.0 | 2.5 | 0.0 | <0.001 | 14.7 | 230.6 | 0.0 | 66.4 | 22.4 | 4.2 | 0.1 | <0.05 |
| 62779 | 7/19/05 | 50 | 11.3 | 7.3 | 461 | 7.67 | 438 | 271.4 | 47.8 | 16.1 | 21.0 | 3.9 | 0.2 | 0.1 | 22.0 | 205.2 | 0.0 | 42.7 | 15.8 | 0.0 | 0.6 | 0.1 |
| 58737 | 7/19/05 | 207 | 11.1 | 7.0 | 549 | 7.41 | 524 | 320.8 | 75.1 | 17.1 | 10.6 | 5.5 | 0.1 | 0.0 | 24.0 | 198.6 | 0.0 | 61.0 | 17.2 | 12.4 | <0.05 | <0.05 |
| 134497 | 7/19/05 | 145 | 16.3 | 7.3 | 661 | 7.61 | 649 | 398.1 | 67.6 | 30.6 | 24.7 | 5.5 | 0.1 | 0.0 | 24.0 | 255.9 | 0.0 | 116.0 | 1.1 | 1.0 | 1.1 | 0.4 |
| 220386 | 7/22/05 | | 11.7 | 7.4 | 411 | 7.78 | 564 | 377.2 | 60.7 | 17.0 | 37.1 | 3.4 | 0.0 | <0.001 | 31.0 | 147.9 | 0.0 | 126.0 | 26.1 | 2.5 | 0.6 | <0.05 |
| 153703 | 7/22/05 | 257 | | | | 8.27 | 377 | 243.5 | 43.2 | 11.4 | 20.6 | 3.0 | <0.005 | <0.001 | 23.5 | 145.4 | 0.0 | 57.1 | 11.4 | 1.3 | 0.4 | <0.05 |
| 182549 | 7/22/05 | 100 | 13.9 | 7.3 | 299 | 7.85 | 379 | 241.1 | 46.2 | 10.0 | 20.7 | 3.2 | <0.005 | <0.001 | 19.9 | 177.9 | 0.0 | 39.5 | 12.9 | 0.8 | 0.3 | <0.05 |
| 60800 | 7/22/05 | 100 | 13.2 | 7.4 | 444 | 7.91 | 445 | 287.0 | 52.9 | 12.7 | 28.3 | 3.3 | <0.005 | <0.001 | 19.9 | 204.7 | 0.0 | 50.8 | 15.6 | 2.3 | 0.3 | <0.05 |
| 134635 | 7/29/05 | 120 | 13.9 | 7.6 | 402 | 8.08 | 486 | 271.1 | 41.5 | 17.2 | 24.8 | 1.2 | 0.0 | <0.001 | 22.7 | 214.5 | 0.0 | 40.8 | 17.0 | 0.0 | 0.2 | <0.05 |
| 130936 | 7/29/05 | 140 | 15 | 7.7 | 379 | 8.04 | 418 | 251.3 | 44.8 | 14.8 | 22.0 | 2.1 | <0.005 | <0.001 | 18.1 | 202.3 | 0.0 | 36.5 | 13.3 | 0.0 | 0.1 | <0.05 |
| 64880 | 7/29/05 | 86 | 11.7 | 7.4 | 574 | 7.95 | 658 | 390.5 | 67.1 | 18.8 | 51.6 | 2.4 | 0.0 | <0.001 | 19.9 | 274.2 | 0.0 | 67.4 | 19.9 | 8.4 | <0.05 | <0.05 |
| 177845 | 8/1/05 | 198 | 10.6 | 7.2 | 524 | 7.53 | 543 | 334.8 | 66.9 | 26.0 | 11.6 | 11.4 | 0.0 | 0.1 | 19.9 | 253.2 | 0.0 | 54.8 | 17.2 | 1.9 | 0.3 | <0.05 |
| 61619 | 8/1/05 | 70 | 11.4 | 6.9 | 338 | 7.74 | 341 | 216.2 | 40.6 | 8.8 | 17.3 | 3.0 | 0.0 | <0.001 | 22.3 | 128.1 | 0.0 | 50.8 | 9.2 | 0.8 | 0.2 | <0.05 |
| 177799 | 8/2/05 | 425 | 19.1 | 7.3 | 746 | 8.01 | 795 | 577.5 | 86.4 | 22.0 | 69.5 | 4.2 | 0.3 | 0.0 | 35.7 | 206.5 | 0.0 | 210.0 | 44.8 | 0.5 | 2.4 | <0.05 |
| | Maximum | | | 7.9 | 1667 | 8.3 | 2580.0 | 1810.8 | 221.0 | 119.0 | 235.0 | 11.4 | 0.3 | 0.1 | 274.0 | 616.0 | 0.0 | 538.0 | 342.0 | 12.4 | 2.4 | 0.7 |
| | Minimum | | | 6.9 | 229 | 7.3 | 273.0 | 199.0 | 29.7 | 6.5 | 10.6 | 0.8 | 0.0 | 0.0 | 9.1 | 128.1 | 0.0 | 29.2 | 1.1 | 0.0 | 0.0 | 0.0 |
| | Median | | | 7.4 | 470 | 7.8 | 543.0 | 333.8 | 58.2 | 21.6 | 24.8 | 3.1 | 0.0 | 0.0 | 22.0 | 214.5 | 0.0 | 62.7 | 17.1 | 2.0 | 0.3 | 0.0 |

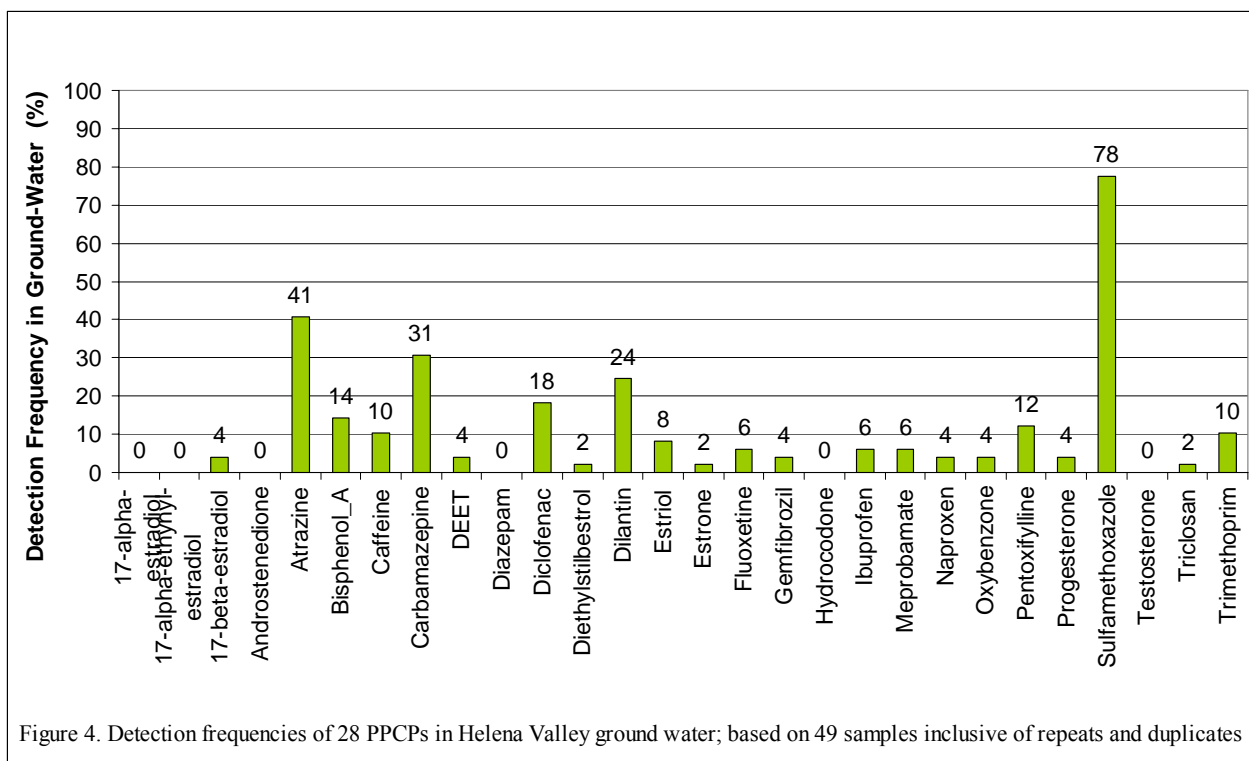
Table 1. Results of dissolved inorganic analyses for 35 well sites in the Helena Valley with maximum, minimum and median values, continued.

| Gwic Id | Sample Date | As (ug/l) | B (ug/l) | Ba (ug/l) | Be (ug/l) | Br (ug/l) | Cd (ug/l) | Co (ug/l) | Cr (ug/l) | Cu (ug/l) | Li (ug/l) | Mo (ug/l) | Ni (ug/l) | Pb (ug/l) | Sb (ug/l) | Se (ug/l) | Sr (ug/l) | Ti (ug/l) | Tl (ug/l) | U (ug/l) | V (ug/l) | Zn (ug/l) | Zr (ug/l) |
|---------|----------------|-----------|----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|----------|----------|-----------|-----------|
| 62523 | 5/24/05 | 0.0 <30 | | 25.8 <2 | | <50 | <1 | <2 | <2 | 4.2 | 6.4 | 10.7 <2 | <2 | <2 | <1 | 385.0 <1 | <5 | | 6.3 <5 | | 2.3 <2 | | |
| 64826 | 5/23/05 | 1.8 | 69.9 | 47.7 <2 | | <50 | <1 | <2 | <2 | <2 | 35.7 <10 | <2 | <2 | <2 | 1.0 | 441.0 <1 | <5 | | 5.9 <5 | | 4.4 <2 | | |
| 5756 | 5/23/05 | 2.7 | 88.9 | 68.0 <2 | | <100 | <1 | <2 | <2 | 2.6 | 19.0 <10 | <2 | <2 | <2 | 3.0 | 654.0 <1 | <5 | | 29.1 | 6.1 | 14.2 <5 | | |
| 62570 | 5/23/05 | 0.0 <150 | | 33.6 <2 | | <100 | <1 | <2 | <10 | <5 | 25.6 <10 | | 3.3 <10 | <10 | 7.8 | 940.0 <1 | <25 | | 17.7 <10 | | 522.0 <2 | | |
| 64806 | 5/23/05 | 2.0 <30 | | 184.0 <2 | | <50 | <1 | <2 | <2 | <2 | 12.8 <10 | <2 | <2 | <2 | 1.5 | 347.0 <1 | <5 | | 4.7 <5 | | 8.5 <2 | | |
| 194850 | 5/24/05 | 2.0 <30 | | 184.0 <2 | | <50 | <1 | <2 | <2 | <2 | 12.8 <10 | <2 | <2 | <2 | 1.5 | 347.0 <1 | <5 | | 4.7 <5 | | 8.5 <2 | | |
| 62369 | 5/31/05 | 3.2 | 188.0 | 35.6 <2 | | <50 | <1 | <2 | 2.2 | 44.7 | 30.3 <10 | <2 | <2 | <2 | 1.7 | 502.0 | 1.1 <5 | | 10.6 | 6.7 | 28.6 <2 | | |
| 62575 | 5/31/05 | 3.2 | 188.0 | 36.4 <2 | | <50 | <1 | <2 | 2.2 | 25.0 | 29.7 <10 | <2 | <2 | <2 | 1.8 | 492.0 <1 | <5 | | 10.7 | 6.8 | 15.4 <2 | | |
| 65388 | 6/5/05 | 1.5 <30 | | 61.3 <2 | | <500 | <1 | <2 | <2 | 2.5 | 17.4 <10 | <2 | <2 | <2 | 4.3 | 519.0 <1 | <5 | | 3.9 <5 | | 14.5 <2 | | |
| 170202 | 6/5/05 | 1.7 | 31.8 | 33.0 <2 | | 109.0 <1 | | <2 | <2 | 3.9 | 48.4 <10 | <2 | | 3.6 <2 | 4.3 | 638.0 <1 | <5 | | 5.5 <5 | | 58.0 <2 | | |
| 187850 | 5/30/05 | 1.6 <30 | | 59.9 <2 | | 203.0 <1 | | <2 | 3.2 <2 | | 16.5 <10 | <2 | <2 | <2 | 4.6 | 509.0 <1 | <5 | | 3.8 <5 | | 7.9 <2 | | |
| 206394 | 5/30/05 | 3.0 <30 | | 99.0 <2 | | 642.0 <1 | | <2 | <2 | 3.4 | 20.9 <10 | <2 | <2 | <2 | 13.3 | 1107.0 | 1.7 <5 | | 4.3 <5 | | 23.1 <2 | | |
| 165085 | 7/15/05 | 1.0 <30 | | 62.6 <2 | | <50 | <1 | <2 | <2 | <2 | 17.0 <10 | <2 | <2 | <2 | 1.3 | 285.0 <1 | <5 | | 2.5 <5 | | 10.5 <2 | | |
| 220274 | 7/14/05 | 7.4 | 37.6 | 24.5 <2 | | 77.0 <1 | | <2 | <2 | 2.3 | 16.2 <10 | <2 | <2 | <2 | 3.5 | 389.0 <1 | <5 | | 7.5 <5 | | 12.1 <2 | | |
| 220272 | 7/15/05 | 2.4 <30 | | 72.5 <2 | | <1 | | <2 | <2 | <2 | 16.0 <10 | <2 | <2 | <2 | 7.5 | 676.0 <1 | <5 | | 4.2 <5 | | 28.3 <2 | | |
| 58685 | 7/15/05 | 5.3 | 31.3 | 21.4 <2 | | <50 | <1 | <2 | <2 | 4.0 | 9.6 <10 | <2 | <2 | <2 | 1.6 | 247.0 <1 | <5 | | 4.5 <5 | | 92.7 <2 | | |
| 58712 | 7/14/05 | 4.0 | 77.7 | 86.0 <2 | | <500 | <1 | <2 | <2 | 13.8 | 15.3 | 13.1 <2 | | 6.3 <2 | 6.1 | 339.0 <1 | <5 | | 11.0 <5 | | 10.5 <2 | | |
| 165017 | 7/19/05 | 1.9 | 90.4 | 45.4 <2 | | <50 | <1 | <2 | <2 | 22.5 | 10.6 <10 | <2 | | 4.5 <2 | <1 | 242.0 <1 | <25 | | 0.0 <25 | | 9.7 <2 | | |
| 65071 | 7/19/05 | 2.7 <30 | | 69.5 <2 | | <50 | <1 | <2 | <2 | 2.1 | 22.6 <10 | <2 | <2 | <2 | <1 | 298.0 <1 | <5 | | 3.3 <5 | <2 | <2 | | |
| 61051 | 7/20/05 | 2.0 <30 | | 61.6 <2 | | <50 | <1 | <2 | <2 | 29.7 | 29.0 <10 | | 3.4 <2 | <2 | <1 | 329.0 <1 | <5 | | 4.0 <5 | | 19.3 <2 | | |
| 61055 | 7/19/05 | 2.1 <30 | | 54.1 <2 | | <50 | <1 | <2 | <2 | <2 | 29.1 <10 | <2 | <2 | <2 | <1 | 295.0 <1 | <5 | | 3.4 <5 | | 2.6 <2 | | |
| 62802 | 7/19/05 | 1.6 <30 | | 29.1 <2 | | <50 | <1 | <2 | 2.1 | 8.5 | 10.2 <10 | <2 | <2 | <2 | 2.0 | 226.0 <1 | <20 | | 2.2 <5 | | 2.4 <2 | | |
| 62779 | 7/19/05 | 17.1 | 59.2 | 33.8 <2 | | <50 | <1 | <2 | <2 | <2 | 47.1 <10 | <2 | <2 | <2 | <1 | 314.0 <1 | <5 | | 6.5 <5 | | 7.4 <2 | | |
| 58737 | 7/19/05 | 1.3 <30 | | 4.0 <2 | | <50 | <1 | <2 | <2 | 8.8 | 12.4 <10 | <2 | | 2.2 <2 | <1 | 329.0 <1 | <5 | | 13.5 | 7.4 | 2.0 <2 | | |
| 134497 | 7/19/05 | 0.0 | 32.0 | 21.6 <2 | | <100 | <1 | <2 | <2 | <2 | 12.4 <10 | <2 | | 2.2 <2 | 3.1 | 333.0 <1 | <5 | | 8.4 <5 | | 2.9 <2 | | |
| 220386 | 7/22/05 | 1.8 | 32.3 | 25.3 <2 | | 178.0 <1 | | <2 | <2 | <2 | 20.0 <10 | <2 | <2 | <2 | 7.1 | 503.0 <1 | <5 | | 7.3 <5 | <2 | <2 | | |
| 153703 | 7/22/05 | 1.8 <30 | | 52.8 <2 | | 83.0 <1 | | <2 | <2 | 6.9 | 15.3 <10 | <2 | <2 | <2 | 3.0 | 391.0 <1 | <5 | | 7.5 <5 | | 2.9 <2 | | |
| 182549 | 7/22/05 | 1.1 | 69.8 | 43.5 <2 | | <50 | <1 | <2 | <2 | <2 | 16.6 | 10.6 <2 | <2 | <2 | 1.1 | 425.0 <1 | <5 | | 10.3 <5 | | 14.5 <2 | | |
| 60800 | 7/22/05 | 1.8 | 62.7 | 51.0 <2 | | <50 | <1 | <2 | <2 | <2 | 18.2 | 18.4 <2 | <2 | <2 | <1 | 494.0 <1 | <5 | | 20.1 <5 | <2 | <2 | | |
| 134635 | 7/29/05 | 2.1 | 90.4 | 59.6 <2 | | <50 | <1 | <2 | <2 | <2 | 18.9 <10 | <2 | <2 | <2 | 1.4 | 272.0 <1 | <5 | | 2.6 <5 | | 8.4 <2 | | |
| 130936 | 7/29/05 | 1.8 | 77.4 | 71.5 <2 | | <50 | <1 | <2 | <2 | 3.0 | 8.0 <10 | <2 | <2 | <2 | <1 | 212.0 <1 | <5 | | 1.8 <5 | <2 | <2 | | |
| 64880 | 7/29/05 | 2.7 | 59.2 | 48.9 <2 | | <50 | <1 | <2 | <2 | 18.4 | 10.3 <10 | <2 | <2 | <2 | 2.4 | 201.0 <1 | <5 | | 2.6 <5 | | 16.5 <2 | | |
| 177845 | 8/1/05 | 1.9 <30 | | 14.5 <2 | | <50 | <1 | <2 | <2 | <2 | 18.6 <10 | <2 | <2 | <2 | 2.4 | 201.0 <1 | <5 | | 2.6 <5 | | 16.5 <2 | | |
| 61619 | 8/1/05 | 1.7 <30 | | 49.5 <2 | | <50 | <1 | <2 | <2 | 22.0 | 16.5 <10 | <2 | <2 | <2 | 4.2 | 327.0 <1 | <5 | | 4.6 <5 | | 89.2 <2 | | |
| 177799 | 8/2/05 | 11.4 | 315.0 | 15.9 <2 | | <50 | <1 | <2 | <2 | <2 | 85.8 <10 | <2 | <2 | <2 | 2.8 | 708.0 <1 | <5 | | 2.2 <5 | | 19.3 <2 | | |
| | Maximum | 17.1 | 315.0 | 184.0 <2 | | 642.0 <1 | | <2 | 3.2 | 44.7 | 85.8 | 18.4 | 3.4 | 6.3 <2 | 13.3 | 1107.0 | 1.7 <5 | | 29.1 | 7.4 | 522.0 <2 | | |
| | Minimum | 0.0 | 31.3 | 4.0 <2 | | 77.0 <1 | | <2 | 2.1 | 2.1 | 6.4 | 10.6 | 3.3 | 2.2 <2 | 1.0 | 201.0 | 1.1 <5 | | 0.0 | 6.1 | 2.0 <2 | | |
| | Median | 1.9 | 69.9 | 48.9 <2 | | 143.5 <1 | | <2 | 2.2 | 6.9 | 17.0 | 11.9 | 3.4 | 3.6 <2 | 2.9 | 347.0 | 1.4 <5 | | 4.7 | 6.7 | 12.1 <2 | | |

Results and Discussion

As shown in figures 4 and 5 and in table 2, out of 28 PPCP analytes sampled from 38 sites, only 6 were not detected in ground water. Detection frequencies, concentrations, and locations of the PPCPs in ground-water samples are shown in figures 4, 5, and 6, respectively. The most frequently detected PPCPs are SMX, atrazine, carbamazepine, dilantin, and diclofenac with detection frequencies of 78, 41, 31, 24, and 18 percent, respectively. Concentrations of SMX range from no detection (ND) to 490 $\eta\text{g/L}$. Maximum concentrations of atrazine, carbamazepine, dilantin, and diclofenac are 130 $\eta\text{g/L}$, 420 $\eta\text{g/L}$, 22 $\eta\text{g/L}$, and 46 $\eta\text{g/L}$, respectively. Table 2 also shows that the occurrence and level of PPCP seem to have little or no correlation to the producing aquifer for each sampled well.

The locations of sites with samples containing SMX, atrazine, and types of wastewater discharges in the Helena valley mapped in figure 7 indicate that although point-sources of wastewater discharges (municipal, storm water, industrial, and a concentrated animal feeding operation) occur within the valley, onsite wastewater discharge appears to dominate water-quality samples. As used in figure 7, the definitions of high and moderate density are greater



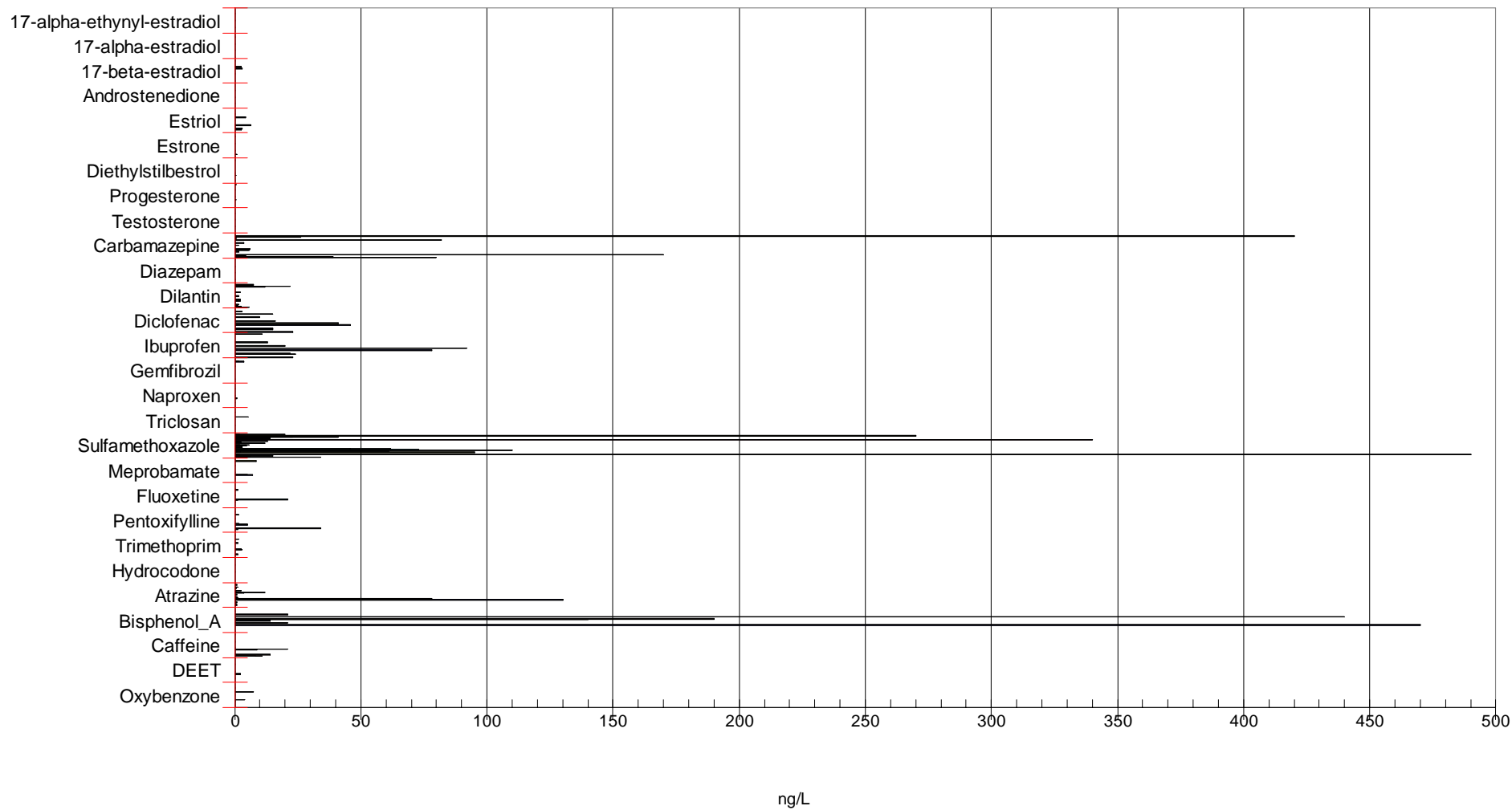
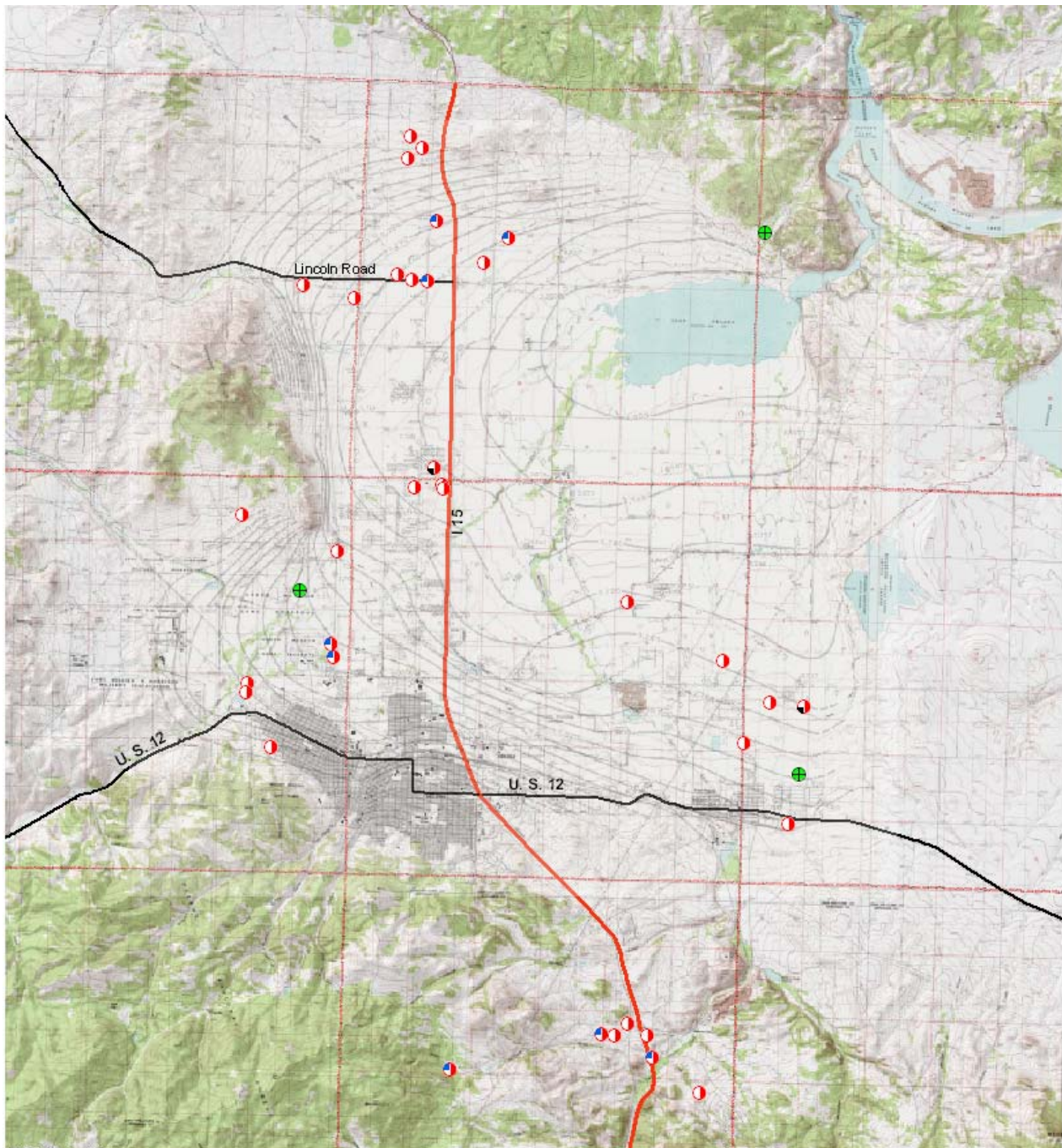


Figure 5. PPCP concentrations.



Explanation:

38 Sites Sampled
 18 PWSs included



Microbial indicators sampled and analyzed for each site:
 Total Coliform (8 sites positive for at least one sampling event)
 Escherichia coli (zero positive samples)
 Enterococci (2 sites positive for at least one sampling event)
 Coliphage - Male Specific and Somatic (zero positive samples)

Endocrine Disrupting and Pharmaceutically Active
 Compounds (EDCs and PACs): Positive detects at 35 of 38 sites



Figure 6. Locations and results of PPCPs and microbial indicators.

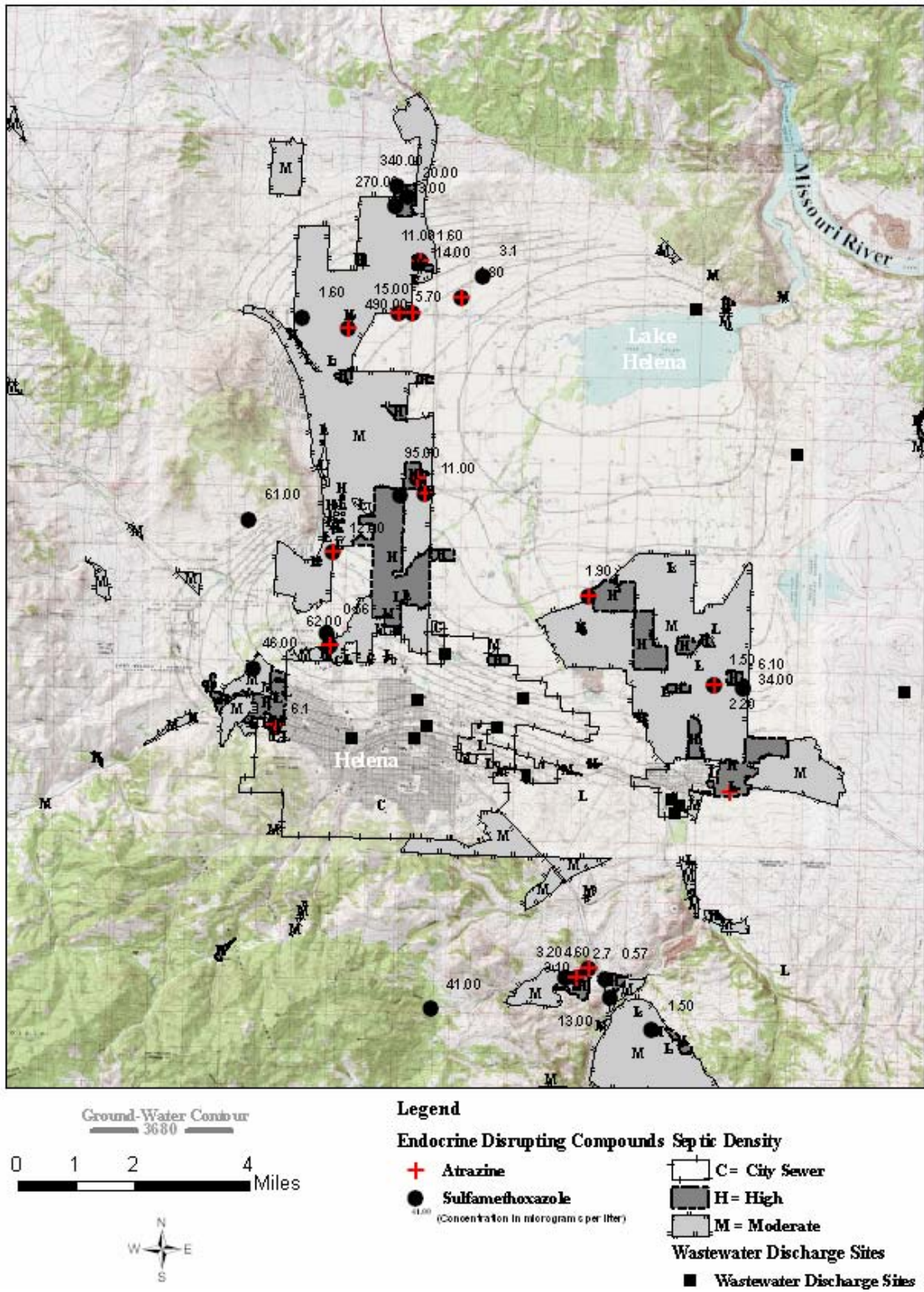


Figure 7. Wastewater discharges, SMX, and atrazine in the Helena valley.

than 300 septic systems (750 persons) per square mile and between 50 (125 persons) and 300 septic systems per square mile, respectively. Figure 8 and table 3 correlate and compare the occurrence of chloride, nitrate, and total dissolved solids (TDS) with SMX and atrazine, the two most frequently detected PPCPs. Chloride, nitrate, and TDS are commonly used inorganic surrogate “indicators” of contamination from domestic wastewater systems. SMX showed no correlation (coefficient < 50%) to chloride, nitrate, or TDS. But atrazine demonstrated 80% correlation with chloride and almost 90% correlation with TDS, suggesting that atrazine may be occurring in domestic wastewater. Atrazine is a triazine herbicide used for the control of broadleaf and grassy weeds, so its presence in domestic wastewater is not expected. SMX is undoubtedly also occurring in domestic wastewater, but it may not be conservative in its flow through septic tanks, perhaps being oxidized by chlorine or other compounds that may be found in wastewater (Dodd, 2004). Complete results of the inorganic analyses can be found at the website, <http://mbmgwic.mtech.edu>.

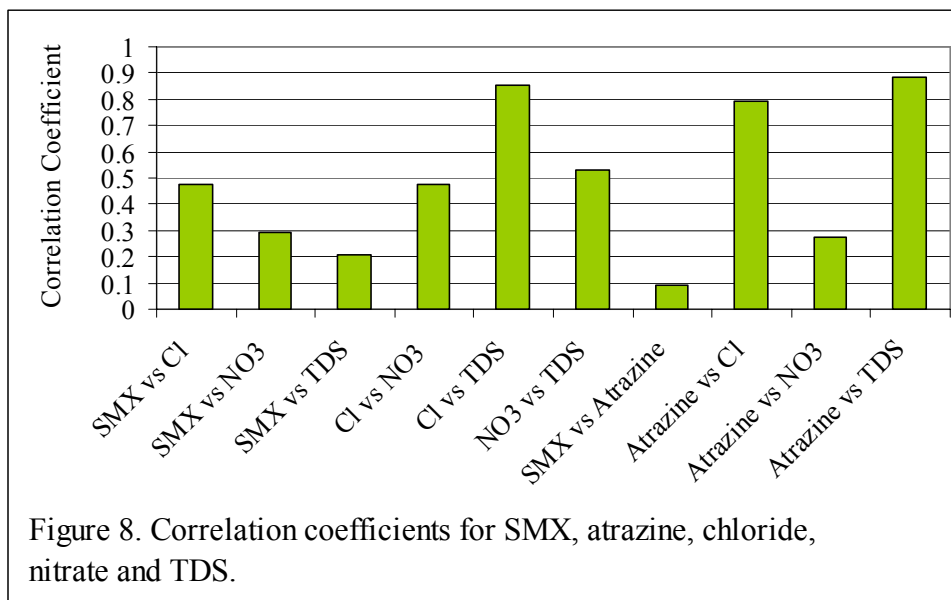


Table 4 presents the results of microbiological analyses of samples collected simultaneously with PPCP samples. As shown in table 4 and figure 6, there were no positive detections of male-specific coliphage, somatic coliphage, or *E. coli* at any of the 38 sites. Yet PPCPs were detected at 32 of the 35 sites. Enterococci were present at 2 different sites at three different times of the year (April, July, and November). Ten positive total coliform samples were detected at 8 different sites. Although the 19 sites used by Kilbreath and others in 2004 were included in the sampling network for this project, his findings were not substantiated in this work. Sample site location within the valley (fig. 6) does not appear to affect the presence or absence of PPCPs or microbial indicators of fecal contamination.

Table 3. Sulfamethoxazole (SMX), atrazine, total dissolved solids (TDS), chloride and nitrate in ground-water samples.

| GWIC ID | SMX (ng/L) | Atrazine (ng/L) | TDS (mg/l) | Cl (mg/l) | NO3 (mg/l) |
|----------------|-------------------|------------------------|-------------------|------------------|-------------------|
| 5756 | 12.00 | 3.50 | 790.2 | 29.5 | 5.4 |
| 58685 | 0.57 | 0.00 | 268.8 | 7.1 | 0.0 |
| 58712 | 3.10 | 1.20 | 511.4 | 35.8 | 10.4 |
| 58737 | 41.00 | 0.00 | 320.8 | 17.2 | 12.4 |
| 60800 | 34.00 | 0.00 | 287.0 | 15.6 | 2.3 |
| 60987 | 0.00 | 0.54 | 377.2 | 26.1 | 2.5 |
| 61051 | 11.00 | 1.30 | 288.8 | 13.4 | 2.0 |
| 61619 | 1.90 | 0.79 | 216.2 | 9.2 | 0.8 |
| 62369 | 61.00 | 0.00 | 546.5 | 17.6 | 9.5 |
| 62523 | 0.00 | 0.00 | 315.2 | 4.8 | 0.0 |
| 62570 | 110.00 | 130.00 | 1810.8 | 342.0 | 9.9 |
| 62575 | 0.56 | 0.00 | 543.3 | 17.6 | 9.4 |
| 62802 | 6.1 | 2.6 | 325.7 | 22.4 | 4.2 |
| 64806 | 3.1 | 0.00 | 528.7 | 17.1 | 8.6 |
| 64826 | 1.80 | 0.91 | 333.8 | 15.1 | 1.9 |
| 64880 | 5.70 | 0.96 | 390.5 | 19.9 | 8.4 |
| 65071 | 95.00 | 0.82 | 280.7 | 16.6 | 3.5 |
| 65388 | 490.00 | 0.59 | 364.0 | 45.2 | 3.9 |
| 134497 | 13.00 | 0.00 | 398.1 | 1.1 | 1.0 |
| 134635 | 14.00 | 0.66 | 271.1 | 17.0 | 0.0 |
| 137172 | 2.00 | 0.00 | 251.3 | 13.3 | 0.0 |
| 153703 | 0.00 | 0.00 | 243.5 | 11.4 | 1.3 |
| 165017 | 15.00 | 0.54 | 288.9 | 15.7 | 1.5 |
| 165085 | 0.00 | 0.00 | 199.0 | 4.9 | 0.9 |
| 170202 | 0.00 | 0.00 | 312.5 | 17.1 | 0.0 |
| 177799 | 2.7 | 12.00 | 577.5 | 44.8 | 0.5 |
| 177845 | 1.50 | 0.00 | 334.8 | 17.2 | 1.9 |
| 182549 | 2.20 | 0.68 | 241.1 | 12.9 | 0.8 |
| 187850 | 20.00 | 0.00 | 364.5 | 43.5 | 3.8 |
| 194850 | 1.60 | 0.00 | 446.3 | 17.6 | 0.7 |
| 206394 | 340.00 | 0.00 | 731.5 | 240.0 | 11.2 |
| 220272 | 270.00 | 0.00 | 425.7 | 96.4 | 4.7 |
| 220274 | 4.60 | 0.00 | 397.4 | 27.1 | 2.0 |

Table 4. Total coliform, *E.coli*, enterococcus and coliphage in ground-water samples.

| GWIC ID | DPHHS Lab ID# | Sample Collection Date | Total Coliform, cfu**/100 ml | E. Coli, cfu/100 ml | Enterococci, cfu/100 ml | Coliphage, Male Specific | Coliphage, Somatic | GWIC ID | DPHHS Lab ID# | Sample Collection Date | Total Coliform, cfu**/100 ml | E. Coli, cfu/100 ml | Enterococci, cfu/100 ml | Coliphage, Male Specific | Coliphage, Somatic |
|---------|---------------|------------------------|------------------------------|---------------------|-------------------------|--------------------------|--------------------|---------|---------------|------------------------|------------------------------|---------------------|-------------------------|--------------------------|--------------------|
| 60800 | W0504-1598 | 04/27/05 | <1 | <1 | 1 | Neg | Neg | 177799 | W0506-2231 | 06/07/05 | <1 | <1 | <1 | Neg | Neg |
| | W0511-4702 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | | W0511-4673 | 11/02/05 | <1 | <1 | <1 | Neg | Neg |
| 177845 | W0511-4670 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | 182549 | W0504-1574 | 04/27/05 | <1 | <1 | <1 | Neg | Neg |
| 165017 | W0507-2958 | 07/19/05 | 4 | <1 | <1 | Neg | Neg | | W0511-4695 | 11/02/05 | <1 | <1 | <1 | Neg | Neg |
| | W0511-4686 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | 134497 | W0507-2963 | 07/19/05 | 16 | <1 | <1 | Neg | Neg |
| 64826 | W0504-1577 | 04/27/05 | <1 | <1 | <1 | Neg | Neg | | W0511-4672 | 11/02/05 | 165 | <1 | <1 | Neg | Neg |
| | W0511-4678 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | 62802 | W0507-2960 | 07/19/05 | <1 | <1 | <1 | Neg | Neg |
| 65388 | W0505-1674 | 05/04/05 | <1 | <1 | <1 | Neg | Neg | | W0511-4850 | 11/09/05 | <1 | <1 | <1 | Neg | Neg |
| | W0511-4687 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | 58685 | W0506-2232 | 06/07/05 | <1 | <1 | <1 | Neg | Neg |
| 60987 | W0504-1597 | 04/27/05 | <1 | <1 | <1 | Neg | Neg | | W0511-4674 | 11/02/05 | <1 | <1 | <1 | Neg | Neg |
| | W0511-4705 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | 206394 | W0505-1673 | 05/04/05 | <1 | <1 | <1 | Neg | Neg |
| | W0504-1582 | 04/27/05 | <1 | <1 | <1 | Neg | Neg | | W-511-4685 | 11/02/05 | <1 | <1 | <1 | Neg | Neg |
| 153703 | W0505-1677 | 05/04/05 | <1 | <1 | <1 | Neg | Neg | 64806 | W0504-1576 | 04/27/05 | <1 | <1 | <1 | Neg | Neg |
| | W0511-4694 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | | W0511-4677 | 11/02/05 | 4 | <1 | <1 | Neg | Neg |
| 65071 | W0507-2959 | 07/19/05 | <1 | <1 | 2 | Neg | Neg | 134632 | W0506-2235 | 06/07/05 | 1 | <1 | <1 | Neg | Neg |
| | W0511-4682 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | | W0511-4680 | 11/02/05 | <1 | <1 | <1 | Neg | Neg |
| 186800 | W0511-4851 | 11/09/05 | <1 | <1 | <1 | Neg | Neg | | W0511-4681 | 11/02/05 | <1 | <1 | <1 | Neg | Neg |
| 194850 | W0504-1580 | 04/27/05 | <1 | <1 | <1 | Neg | Neg | 61051 | W0507-2955 | 07/19/05 | <1 | <1 | <1 | Neg | Neg |
| | W0511-4684 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | | W0511-4704 | 11/02/05 | <1 | <1 | <1 | Neg | Neg |
| 62369 | W0505-1675 | 05/04/05 | <1 | <1 | <1 | Neg | Neg | 61055 | W0507-2956 | 07/19/05 | <1 | <1 | <1 | Neg | Neg |
| | W0511-4700 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | 58737 | W0507-2962 | 07/19/05 | 4 | <1 | <1 | Neg | Neg |
| 62575 | W0505-1676 | 05/04/05 | 9 | <1 | <1 | Neg | Neg | | W0511-4671 | 11/02/05 | <1 | <1 | <1 | Neg | Neg |
| | W0511-4698 | 11/02/05 | 12 | <1 | <1 | Neg | Neg | 220272 | W0506-2230 | 06/08/05 | <1 | <1 | <1 | Neg | Neg |
| 62570 | W0504-1601 | 04/27/05 | 1 | <1 | <1 | Neg | Neg | | W0511-4683 | 11/02/05 | <1 | <1 | <1 | Neg | Neg |
| | W0511-4696 | 11/02/05 | <1 | <1 | <1* | Neg | Neg | 62779 | W0507-2961 | 07/19/05 | <1 | <1 | <1 | Neg | Neg |
| | W0511-4697 | 11/02/05 | <1 | <1 | 1* | Neg | Neg | 62523 | W0504-1603 | 04/27/05 | <1 | <1 | <1 | Neg | Neg |
| 62597 | W0511-4849 | 11/09/05 | <1 | <1 | <1 | Neg | Neg | | W0511-4701 | 11/02/05 | <1 | <1 | <1 | Neg | Neg |
| 58712 | W0506-2233 | 06/07/05 | <1 | <1 | <1 | Neg | Neg | 61619 | W0504-1600 | 04/27/05 | <1 | <1 | <1 | Neg | Neg |
| | W0511-4675 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | | W0511-4703 | 11/02/05 | <1 | <1 | <1 | Neg | Neg |
| | W0511-4676 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | 187850 | W0504-1578 | 04/27/05 | <1 | <1 | <1 | Neg | Neg |
| 170202 | W0504-1575 | 04/27/05 | <1 | <1 | <1 | Neg | Neg | | W0511-4689 | 11/02/05 | <1 | <1 | <1 | Neg | Neg |
| | W0511-4679 | 11/01/05 | <1 | <1 | <1 | Neg | Neg | | | | | | | | |
| | W0511-4691 | 11/01/05 | <1 | <1 | <1 | Neg | Neg | | | | | | | | |
| 220274 | W0506-2234 | 06/07/05 | <1 | <1 | <1 | Neg | Neg | | | | | | | | |
| | W0511-4669 | 11/02/05 | 2 | <1 | <1 | Neg | Neg | | | | | | | | |
| 64880 | W0504-1579 | 04/27/05 | <1 | <1 | <1 | Neg | Neg | | | | | | | | |
| | W0511-4688 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | | | | | | | | |
| 165085 | W0507-2957 | 07/19/05 | <1 | <1 | <1 | Neg | Neg | | | | | | | | |
| | W-511-4692 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | | | | | | | | |
| | W0511-4693 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | | | | | | | | |
| 137172 | W0504-1581 | 04/27/05 | <1 | <1 | <1 | Neg | Neg | | | | | | | | |
| | W0511-4690 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | | | | | | | | |
| 160324 | W0504-1599 | 04/27/05 | <1 | <1 | <1 | Neg | Neg | | | | | | | | |
| 5756 | W0504-1602 | 04/27/05 | <1 | <1 | <1 | Neg | Neg | | | | | | | | |
| | W0511-4699 | 11/02/05 | <1 | <1 | <1 | Neg | Neg | | | | | | | | |

* Growth in Enterococcus media

** cfu = colony forming units

The lack of positive coliphage detections in the presence of PPCPs point to its unsuitability as an indicator of fecal contamination in ground water. Whether coliphage is being attenuated in the subsurface as suggested by the USGS (2005) or whether the poor reproducibility of results is attributable to laboratory or sampling error, the argument can be made that coliphage results are difficult to reproduce in the field, casting its utility as an indicator organism into question. Based on the data in table 4, it appears that of the five microorganisms, total coliform is the most reliable indicator of fecal contamination. Both *E. coli* and enterococci are associated with fresh sewage. A drawback to using these two indicators in regional ground-water settings is that they may die out more quickly or be less mobile in the subsurface than some waterborne pathogens, thereby rendering them even less suitable as indicator organisms than coliphage.

Conclusions

The Helena Valley in west-central Montana is experiencing rapid growth into previously un-sewered areas that rely on septic tanks and drainfields for onsite wastewater treatment and disposal. Detections of PPCPs in drinking water derived from wells is consistent with the findings of other investigators who are evaluating the occurrence of these compounds in ground-water and septic systems.

SMX and atrazine, the two most frequently detected compounds, were found at frequencies of 80% and 40% of samples, respectively. A comparison of SMX and atrazine with chloride, TDS, and nitrate shows that atrazine demonstrates a strong correlation with chloride and TDS, two typical inorganic indicators of ground-water degradation from domestic wastewater. Further sampling and analysis of septic tank effluent should be conducted to verify whether atrazine is occurring in domestic wastewater.

While there are limited detections of total coliform and enterococci, PPCPs are consistently detected in the absence of both male-specific and somatic coliphage as well as *E. coli*. These results present implications for the suitability of coliphage and *E. coli* as indicators of fecal contamination in ground water. Total coliform, though detected at only eight sites, was superior to coliphage as an indicator organism in this ground-water setting.

The human and aquatic effects from chronic exposure and ingestion of PPCPs at $\mu\text{g/L}$ or sub- $\eta\text{g/L}$ concentrations are mostly unknown as are potential synergistic or additive effects of exposure and ingestion of PPCP mixtures such as those found in the Helena valley. Since the ground water ultimately discharges to the Missouri River, it is hoped that effects on human health and aquatic ecosystems become better understood.

Future investigations should include fate and transport studies that evaluate the role of various aquifer properties and water-quality parameters in controlling PPCP sorption and degradation and coliphage survival and attenuation in the subsurface. PPCP fate in advanced onsite wastewater treatment systems should also be further evaluated.

Acknowledgments

This work was funded by the Montana Water Center and is a collaborative effort between the Montana Department of Environmental Quality, the Lewis and Clark County Water Quality Protection District, the Montana Department of Public Health and Human Services, and the Montana Bureau of Mines and Geology. Thanks go to DEQ staff Carolyn DeMartino, Jeffrey Herrick, Jim Stimson, and Eric Sivers for field support. Special recognition goes to Jim Stimson for GIS support and map production.

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