

Available online at www.sciencedirect.com



Environmental and Experimental Botany

Environmental and Experimental Botany 62 (2008) 290-299

www.elsevier.com/locate/envexpbot

Responses of two poplar species (*Populus alba* and *Populus x canadensis*) to high copper concentrations

M. Borghi^{a,*}, R. Tognetti^b, G. Monteforti^a, L. Sebastiani^a

^a Scuola Superiore Sant' Anna, Piazza Martiri

della Libertà 33, I-56127 Pisa, Italy

^b Università degli Studi del Molise, Dipartimento di Scienze e Tecnologie per l'Ambiente e il Territorio (STAT),

Contrada Fonte Lappone, I-86090 Pesche (IS), Italy

Received 17 January 2007; received in revised form 12 July 2007; accepted 1 October 2007

Abstract

With the aim to examine their potentials as renewable resources to decontaminate polluted soils, growth, photosynthesis and nitrogen balance were analyzed in two poplar species (*Populus x canadensis*, *Adda* clone and *Populus alba*, *Villafranca* clone) to investigate the tolerance to high copper (Cu) concentrations. The two clones showed different responses to Cu in terms of tolerance and metal allocation: *P. x canadensis* accumulated Cu in roots, displaying features sought in plants suitable for phytostabilization, while *P. alba* accumulated the metal in leaves, like an indicator species.

© 2007 Elsevier B.V. All rights reserved.

Keywords: Poplar; Heavy metals; Gas exchange; Phytotoxicity

1. Introduction

Pollution of soil and water with heavy metals has serious consequences for environment and human health. In particular, human health is affected through direct exposure via skin and inhalation, and through oral exposure, with potential for systemic accumulation by consumption of contaminated food and groundwater (Hrudey et al., 1995).

Plants show different degrees of tolerance to heavy metals, being able to exclude or accumulate and store them in particular organs, tissues, modified cells (trichomes) and subcellular compartments (vacuoles). Several species are known to accumulate large amounts of heavy metal and are therefore called hyperaccumulators. This ability can be used in phytoextraction technologies to bring the level of soil pollution below the limits established by environmental rules (Salt et al., 1998). Responses to high levels of heavy metals have been extensively studied in hyperaccumulators, in model plants, and to a lesser extent

Tel.: +39 050 883 000; fax: +39 050 883 495.

also in trees (Lasat, 2002; Maegher, 2000; Pulford and Watson, 2003).

In trees, levels of tolerance have been proved for many contaminants, so that together with morphological and physiological characteristics (fast growth, deep root system, high hydraulic control capacity and long life) and commercial uses (paper and wood industry), trees appear suitable for phytoremediation purposes (Fischerová et al., 2006; Rokwood et al., 2004).

However, tolerance characteristics and metal uptake and compartmentalization are highly variable among tree species and hybrids (Kopponen et al., 2001; Laureysens et al., 2005; Sebastiani et al., 2004). Responses of trees to high heavy metal concentrations have been studied on polluted fields like mines and disposal sites (Álvarez et al., 2003; Djingova et al., 1999; Madejón et al., 2004), as well as in controlled environments particularly suitable for preliminary screenings (Di Baccio et al., 2003; Punshon et al., 1995).

In this context, the aim of this work was to determine the ability of poplar trees to tolerate high copper (Cu) concentrations and to accumulate Cu ions in aerial plant organs or conversely to store them in the root system. Since metal tolerance and translocation capacity are traits highly dependent on species and clones, we performed our investigation on two different poplar

^{*} Corresponding author at: Scuola Superiore Sant' Anna, BioLabs, Polo Sant' Anna-Valdera, Viale R. Piaggio 34, I-56025 Pontedera (PI), Italy.

E-mail address: m.borghi@sssup.it (M. Borghi).

^{0098-8472/\$ -} see front matter © 2007 Elsevier B.V. All rights reserved. doi:10.1016/j.envexpbot.2007.10.001

species: the white poplar *Populus alba*, *Villafranca* clone and the hybrid *Populus x canadensis*, *Adda* clone. The main physiological responses, with particular emphasis to those mostly disturbed by heavy metals (growth, photosynthesis and nitrogen balance), were discussed.

2. Materials and methods

2.1. Plant material and treatments

Woody cuttings of P. x canadensis, Adda clone and P. alba, Villafranca clone were rooted and grown in pots containing a universal loam:perlite mixture (1:1, v:v) and irrigated daily with deionized water for about 30 days. Plants were successively transferred to a growth chamber (temperature 23 °C; relative humidity 55-75% day-night; photoperiod 16-8h light-dark; maximum light intensity $400-500 \,\mu\text{mol}\,\text{m}^{-2}\,\text{s}^{-1}$), grown in pots filled with clay suitable for hydroponic culture (diameter 13-15 mm; pH 7.0; AgriLeca, Milano, Italia) and supplied with Hoagland's solution, pH 6.7 (Arnon and Hoagland, 1940). After 15 days of acclimation to the hydroponic system, 24 homogeneous plants per genotype were selected and randomly assigned to three groups of treatment with Hoagland's solution containing different concentrations of CuSO₄·5H₂O: 0.4 (control), 25 and 75 μ M and monitored for 30 (T_1) and 80 (T_2) days. To avoid root necrosis the solution was completely replaced twice per week and aerated by diffused air aeration systems. Four plants randomly selected at the beginning of the experiment were used to determine the initial biomass (T_0) after oven drying at 70 °C until constant weight.

2.2. Growth analysis

At times T_1 and T_2 of sampling, four plants from each treatment group were randomly selected and divided into leaves, stem and roots, separately weighted and oven-dried at 70 °C until constant weight was reached. To remove Cu ions adsorbed at the root surface, root apparatus was carefully rinsed with deionised water soon after harvesting. Stem length (SL) and stem basal diameter (SD) were measured weekly on each plant. Leaf area (LA) was determined by the regression equations y=0.7121x ($R^2=0.99$) for P. x canadensis and y=0.6327x ($R^2=0.94$) for P. alba. Both equations were forced through the origin; (y) represents leaf area and (x)the product of the two main orthogonal diameters of leaves. Each equation was previously determined by measuring the surface area of 215 (P. x canadensis) and 184 (P. alba) randomly selected leaves and analyzed using the software Scion Image 4.0.2 (Scion Corporation, MD, USA). Specific leaf area (SLA) was calculated as the ratio of leaf area to corresponding leaf dry mass. Relative growth rate (RGR) was calculated as RGR = $\left[\left(\log_e MT_x - \log_e MT_0 \right) / (T_x - T_0) \right]$ where MT_x is the total dry mass at sampling time T_x (T_1 and T_2) and MT₀ is the initial (T_0) total plant dry mass average (Hunt, 1978). Metal resistance was calculated as tolerance index (TI; %) accordingly to Landberg and Greger (2002) on the basis of total dry weight of shoot and root.

2.3. Copper and nitrogen determinations

Samples of leaves, stem and roots were oven-dried at 70 $^{\circ}$ C and ground to powder in a laboratory mill. The plant material (0.2 g) was mineralized in HNO₃–HClO₄ (5:1, v:v), clarified with ultra pure water and used for measurements of total Cu in an atomic absorption spectrophotometer (Analyst 200, Perkin-Elmer, Norwalk, CT, USA).

Organic nitrogen content of leaves was determined with the Kijeldhal method using a 2100 Kjeltec System Distillation Unit (Foss Tecator AB, Höganäs, Sweden).

2.4. Chlorophyll analysis

Chlorophyll *a* and *b* (Chl_{*a*} and Chl_{*b*}) contents were determined by reading the absorbance at 664 and 647 nm of extracts obtained from two disks of 10 mm in diameter from the fourth to fifth leaf from the apex. The extraction was performed with *N*,*N*-dimethylformamide (1:20, w:v) for 48 h at 4 °C in the dark and chlorophyll content determined according to Moran (1982). Spectrophotometer measurements were monitored on a Lambda 25 Spectrophotometer (Perkin-Elmer, Norwalk, CT, USA).

2.5. Gas exchange and chlorophyll fluorescence measurements

Gas exchange was measured with a portable photosynthesis system (LI-6400, Li-Cor, Lincoln, NE, USA) equipped with CO₂ control module and light source consisting of blue-red light-emitting diodes (LI-6400-02B). Net photosynthesis (A), stomatal conductance to $H_2O(g_s)$ and intercellular CO₂ concentration (C_i) were measured on the fourth and fifth fully expanded leaves from the apex (Leaf Plastochron Index 4-5) (Dickmann, 1971). All gas exchange measurements were conducted at time T_2 on four replicate plants for each clone-treatment combination. Leaf chamber temperature and humidity were adjusted to maintain a leaf-to-air vapour pressure difference of about 1.2 kPa, with irradiance maintained at a previously determined saturating value of 500 μ mol m⁻² s⁻¹. Photosynthetic CO₂ response curves (A/C_i) were determined by measuring the response of A to varying C_i . External CO₂ partial pressures (C_a) were supplied in steps; the CO₂ assimilation rate was first measured by setting the reference CO₂ concentration near ambient (40 Pa) and then at 30, 20, 10, 5, 40, 40, 60 and 80 Pa. Measurements were recorded automatically at each C_a set point when A had equilibrated, which was typically less than 3 min. The A/C_i curves were analyzed with Photosynthesis Assistant software (Version 1.1, Dundee Scientific, Dundee, Scotland). Photosynthesis Assistant analyses A/C_i curves were based on the model of Farquhar et al. (1980), as subsequently modified by von Caemmerer and Farquhar (1981), Sharkey (1985), and Harley and Sharkey (1991). A/C_i curves were analyzed to estimate two biochemical parameters potentially limiting to photosynthesis: V_{cmax} (maximum rate of carboxylation) and J_{max} (light-saturated rate of electron transport). We analyzed the A/C_i curves using the Michaelis-Menten constants of Rubisco described in Harley et al. (1992a,b) and used by Wullschleger (1993). The specificity factor for Rubisco (Jordan and Ogren, 1984) was also analyzed. The relative stomatal limitation to photosynthesis (l_g) , an estimate of the proportion of the reduction in photosynthesis attributable to CO₂ diffusion between the atmosphere and the site of carboxylation, was calculated from A/C_i curves by the method of Farquhar and Sharkey (1982) as $l_g = (1 - A/A_0) \times 100$ where A is the rate of net photosynthesis at the growth C_a and A_o is the photosynthetic rate when C_i equals the growth C_a . Under these conditions, A_o is the rate of photosynthesis that would occur if there were no diffusive limitations to CO₂ transfer from the bulk atmosphere to the site of carboxylation. For this calculation, mesophyll conductance was considered to be infinitely large.

Contemporary to gas exchange determinations, chlorophyll fluorescence emissions in 30-min dark-adapted leaves were measured with a portable pulse amplitude modulation fluorometer (PAM-2000, Walz, Effeltrich, Germany) to estimate the effect of treatments on photosystem II (PSII) efficiency. Background fluorescence signal (F_o), the maximum fluorescence (F_m), and potential quantum yield of PSII photochemistry [$F_v/F_m = (F_m - F_o)/F_m$] were determined as well.

2.6. Statistical analysis

The experiment had a completely randomized design with four replicates per clone for each time point. Data were subjected to analysis of variance (ANOVA) to examine the effect of time, treatment and clone. Statistical analysis was conducted using the software CoStat 6.2 (CoHort Software, CA, USA). Separation of means was performed using LSD test at P = 0.05 significance level.

3. Results

3.1. Growth traits

Increase in the biomass of all plant organs (leaves, stems and roots) from time T_1 to time T_2 was observed in both *P. x canadensis*, *Adda* clone and *P. alba*, *Villafranca* clone under all Cu concentrations tested (Table 1). At time T_2 and 75 μ M Cu, *P. x canadensis* showed a 22% reduction of total biomass compared to control plants, while no statistically significant differences were observed at 25 μ M Cu. In *P. alba* at 25 and 75 μ M Cu, growth was, respectively, 57 and 73% less than in control plants.

A similar trend was observed in leaf growth (leaf area, LA; specific leaf area, SLA), stem growth (stem length, SL; stem diameter, SD), and total growth (relative growth rate, RGR), as shown in Table 2. Therefore, *P. x canadensis, Adda* clone appeared to be less responsive than *P. alba, Villfranca* clone to Cu treatments.

Table 1

Leaves, stems, roots and total biomass of *Populus x canadensis*, *Adda* clone and *Populus alba*, *Villafranca* clone grown with 0.4 (control), 25 and 75 μ M Cu for 30 (T_1) or 80 (T_2) days

Time	Organ	Species	Cu treatment				
				0.4 µM	25 μΜ		75 µM
$\overline{T_1}$	Leaves	P. x canadensis		5.630 ± 0.564	6.341 ±	0.779	3.854 ± 0.790
		P. alba		3.336 ± 0.547	$2.973 \pm$	0.377	2.104 ± 0.584
	Stem	P. x canadensis		2.046 ± 0.258	$2.661 \pm$	0.389	1.464 ± 0.364
		P. alba		1.214 ± 0.248	$1.182 \pm$	0.155	0.835 ± 0.230
	Roots	P. x canadensis		0.804 ± 0.105	$0.627 \pm$	0.104	0.363 ± 0.104
		P. alba		0.398 ± 0.099	$0.383 \pm$	0.044	0.409 ± 0.092
	Total	P. x canadensis		8.484 ± 0.902	9.629 ±	1.251	5.680 ± 1.247
		P. alba		4.948 ± 0.872	4.537 ±	0.552	3.348 ± 0.903
T_2	Leaves	P. x canadensis		22.130 ± 2.169	22.369 ±	3.345	18.336 ± 3.345
		P. alba		27.701 ± 3.486	$12.127 \pm$	2.983	7.291 ± 1.831
	Stem	P. x canadensis		15.530 ± 0.453	11.918 ±	1.654	9.257 ± 1.655
		P. alba		16.542 ± 2.317	5.913 ±	1.178	4.190 ± 0.999
	Roots	P. x canadensis		3.381 ± 0.348	$3.884 \pm$	0.885	4.103 ± 0.759
		P. alba		3.593 ± 0.685	$2.387 \pm$	0.519	1.111 ± 0.198
	Total	P. x canadensis		41.040 ± 1.933	$38.170 \pm$	5.812	31.696 ± 5.789
		P. alba		47.835 ± 6.336	20.427 ±	4.595	12.592 ± 2.995
Organ	Main effects			Interactions			
	Clone	Metal	Time	$\overline{C \times M}$	$\mathbf{C} \times \mathbf{T}$	$M \times T$	$C \times M \times T$
ANOVA P valu	es						
Leaves	.0034	.0003	.0000	.0123	ns	.0035	.0148
Stem	.0013	.0000	.0000	.0421	ns	.0000	ns
Roots	.0028	ns	.0000	ns	.0209	ns	.0206
Total	.0002	.0020	.0000	ns	.0000	ns	ns

Values are the mean \pm standard error (*n*=4). Units are g of dry mass. ANOVA: *P* values for main effects and interactions clone \times metal (C \times M), clone \times time (C \times T), metal \times time (M \times T), clone \times metal \times time (C \times M \times T) are shown.

Table 2

Time	Parameter	Species		Cu treatment			
				0.4 µM	25 µM		75 µM
<i>T</i> ₁	SL	P. x canadensis		0.531 ± 0.030	0.596 ± 0.022		0.425 ± 0.044
		P. alba		0.404 ± 0.035	0.394 :	± 0.022	0.285 ± 0.041
	SD	P. x canadensis		6.975 ± 0.184	7.250 ± 0.433		6.325 ± 0.506
		P. alba		5.000 ± 0.000	5.500 :	± 0.000	4.500 ± 0.354
	LA	P. x canad	lensis	0.157 ± 0.012	0.174 ± 0.022		0.098 ± 0.017
		P. alba		0.105 ± 0.014	0.092 :	± 0.012	0.051 ± 0.015
	SLA	P. x canad	lensis	0.028 ± 0.001	0.027 ± 0.001		0.026 ± 0.001
		P. alba		0.032 ± 0.002	0.031 ± 0.001		0.024 ± 0.000
	RGR	P. x canadensis		0.089 ± 0.004	0.085 ± 0.005		0.066 ± 0.007
		P. alba		0.094 ± 0.006	0.078 ± 0.004		0.066 ± 0.009
T_2	SL	P. x canadensis		1.123 ± 0.039	0.958 ± 0.052		0.819 ± 0.065
		P. alba		1.027 ± 0.005	0.789 ± 0.085		0.755 ± 0.088
	SD	P. x canad	lensis	11.945 ± 0.419	9.860 :	± 0.326	9.710 ± 0.785
		P. alba		12.183 ± 0.848	7.673 :	± 0.393	7.140 ± 0.902
	LA	P. x canadensis		0.433 ± 0.050	0.380 ± 0.046		0.291 ± 0.049
		P. alba		0.579 ± 0.043	0.239 :	± 0.052	0.101 ± 0.044
	SLA	P. x canad	lensis	0.020 ± 0.002	0.017 ± 0.001		0.016 ± 0.001
		P. alba		0.021 ± 0.002	0.020 :	± 0.001	0.015 ± 0.005
	RGR	P. x canadensis		0.050 ± 0.001	0.049 ± 0.002		0.046 ± 0.003
		P. alba		0.059 ± 0.002	0.047	± 0.003	0.041 ± 0.003
Parameter	Main effects			Interactions			
	Clone	Metal	Time	$\overline{C \times M}$	$\mathbf{C} \times \mathbf{T}$	$M \times T$	$C \times M \times T$
ANOVA P valu	es						
SL	.0001	.0000	.0000	ns	ns	.0068	ns
SD	.0000	.0000	.0000	ns	ns	.0001	ns
LA	.0049	.0000	.0000	.0027	ns	.0000	.0044
SLA	.1618	.0016	.0000	ns	ns	ns	ns
RGR	ns	.0007	.0000	ns	ns	ns	ns

Stem length (SL), stem diameter (SD), leaf area (LA), specific leaf area (SLA), relative growth rate (RGR) of *P. x canadensis*, *Adda* clone and *P. alba*, *Villafranca* clone grown with 0.4 (control), 25 and 75 μ M Cu for 30 (T_1) or 80 (T_2) days

Values are mean \pm standard error (n = 4). Units are m (SL), mm (SD), m² (LA), m² g⁻¹ (SLA) and days⁻¹ (RGR). ANOVA: *P* values for main effects and interactions clone \times metal (C \times M), clone \times time (C \times T), metal \times time (M \times T), clone \times metal \times time (C \times M \times T) are shown.

3.2. Nitrogen and chlorophyll concentrations

Nitrogen content in leaves decreased through the treatments in both species. At time T_2 , it decreased more in *P. x canadensis* than in *P. alba*, which showed a slight decrease only in the plants treated with 75 μ M Cu (Table 3).

The two poplar species revealed significant differences in chlorophyll content in the response to Cu treatment (Table 4; see *P* values for metal and interaction). *P. x canadensis* showed a slight increase in Chl_a and Chl_b contents in treated plants compared to controls. In *P. alba* Chl_a and Chl_b contents slightly decreased in plants treated with 75 μ M Cu. Chl_{a/b} ratio did not change through the treatments in both species, indicating that the synthesis of photosystem cores and of pigments of light harvesting complexes was not affected.

3.3. Photosynthetic capacity

Photosynthetic rates maintained stable across Cu treatments in *P. x canadensis* (within 17–14% positive variation compared with control plants), while showed a decrease (about 28%) with increasing Cu concentration in *P. alba* (Fig. 1). Accordingly,

Table 3

N concentration (N%) in leaves of *P. x canadensis*, *Adda* clone and *P. alba*, *Villafranca* clone grown with 0.4 (control), 25 and 75 μ M Cu for 30 (T_1) or 80 (T_2) days

Time	Species		Cu treatment				
			0.4 µM	25 µ.	М	75 μΜ	
T_1	P. x canadensis P. alba		$\begin{array}{c} 3.61 \pm 0.10 \\ 3.35 \pm 0.20 \end{array}$	3.51 3.33	$\pm 0.11 \\ \pm 0.09$	3.11 ± 0.12 2.83 ± 0.09	
<i>T</i> ₂	P. x canadensis P. alba		$\begin{array}{ll} 3.10 \pm 0.11 & 2.44 \pm 0. \\ 2.71 \pm 0.08 & 2.72 \pm 0. \end{array}$		± 0.12 ± 0.03	$\begin{array}{c} 2.11 \pm 0.11 \\ 2.61 \pm 0.11 \end{array}$	
Main ef	fects		Interaction	18			
Clone	Metal	Time	$\overline{C \times M}$	$\mathbf{C} imes \mathbf{T}$	$M \times T$	$C \times M \times T$	
ANOVA ns	A <i>P</i> values .0000	.0001	.0118	.0034	ns	.0351	

Values are mean \pm standard error (*n*=4). Units are expressed as %, w:w (N). ANOVA: *P* values for main effects and interactions clone × metal (C × M), clone × time (C × T), metal × time (M × T), clone × metal × time (C × M × T) are shown. Table 4

294

Chlorophyll *a* (Chl_{*a*}), *b* (Chl_{*b*}), total (Chl_{tot}) and chlorophyll *a/b* ratio (Chl_{*a/b}) in leaves of P. x canadensis, Adda clone and P. alba, Villafranca clone grown with 0.4 (control), 25 and 75 \muM Cu for 80 (T_2) days</sub>*

Time	Chl	Species	Cu treatment			
			0.4 µM	25 µM	75 μΜ	
$\overline{T_2}$	Chl_a	P. x canadensis	8.436 ± 0.672	13.056 ± 0.592	10.248 ± 0.729	
		P. alba	10.200 ± 0.475	10.323 ± 0.676	9.101 ± 0.490	
	Chl_b	P. x canadensis	2.371 ± 0.158	3.544 ± 0.165	3.003 ± 0.216	
		P. alba	2.883 ± 0.131	3.001 ± 0.196	2.658 ± 0.166	
	Chl _{a/b}	P. x canadensis	3.530 ± 0.061	3.686 ± 0.041	3.443 ± 0.124	
		P. alba	3.540 ± 0.037	3.440 ± 0.018	3.438 ± 0.031	
	Chl _{tot}	P. x canadensis	10.807 ± 0.829	16.600 ± 0.754	13.251 ± 0.922	
		P. alba	13.083 ± 0.604	13.324 ± 0.872	11.759 ± 0.656	
Chl		Main effects			Interaction	
		Clone		Metal	$\mathbf{C} \times \mathbf{M}$	
ANOVA P val	ues					
Chl _a		ns		.0006	.0023	
Chl_b		ns		.0019	.0092	
Chl _{a/b}		ns		ns	.0808	
Chltot		ns		.0007	.0030	

Values are mean \pm standard error (n = 4). Units are expressed as μ g mm⁻². ANOVA: *P* values for main effects and interaction clone × metal (C × M) are shown.

stomatal conductance decreased in *P. x canadensis* (44 and 25% negative variation in 25 and 75 μ M treated plants, respectively), while remained fairly steady in *P. alba* (within 2–15% negative variation). Susceptibility for stomata closure with increasing Cu treatment was confirmed in *P. x canadensis* by the decrease of intercellular CO₂ concentration, which was not the case in *P. alba*. However, instantaneous gas exchange parameters did not change significantly with treatment or genotype. Maximum

rate of carboxylation (V_{cmax}) and light-saturated rate of electron transport (J_{max}) showed an increase in 25 μ M Cu treated plants of *P. x canadensis*; whereas, in *P. alba* values of both parameters progressively decreased from control to 75 μ M Cu (Table 5). The J_{max}/V_{cmax} ratio remained steady across treatments in *P. x canadensis* ranging about 2.5, while it decreased in *P. alba* with increasing Cu because of the relatively stronger decrease in J_{max} compared with V_{cmax} . Relative stomatal limitation to photosyn-

Table 5

Maximum rate of carboxylation (V_{cmax}), light-saturated rate of electron transport (J_{max}), relative stomatal gas-phase limitation (l_g) estimated from A/C_i curves and chlorophyll fluorescence in leaves of *P. x canadensis*, *Adda* clone and *P. alba*, *Villafranca* clone grown with 0.4 (control), 25 and 75 μ M Cu for 80 (T_2) days

Time	Parameter	Species	Cu treatment			
			0.4 µM	25 μΜ	75 μΜ	
$\overline{T_2}$	V _{cmax}	P. x canadensis	27.67 ± 2.01	42.73 ± 2.66	35.37 ± 5.57	
		P. alba	45.33 ± 4.97	41.77 ± 4.97	36.63 ± 3.01	
	J_{\max}	P. x canadensis	69.90 ± 2.29	106.63 ± 7.34	89.23 ± 11.87	
		P. alba	153.00 ± 14.08	111.93 ± 14.01	75.60 ± 26.47	
	$J_{\rm max}/V_{\rm cmax}$	P. x canadensis	2.54 ± 0.11	2.52 ± 0.28	2.55 ± 0.16	
		P. alba	3.38 ± 0.25	2.65 ± 0.30	2.07 ± 0.42	
	lo	P. x canadensis	10.08 ± 1.60	22.52 ± 2.79	32.17 ± 9.74	
	5	P. alba	26.58 ± 8.38	22.26 ± 3.48	15.22 ± 7.60	
	$F_{\rm v}/F_{\rm m}$	P. x canadensis	0.75 ± 0.08	0.82 ± 0.00	0.82 ± 0.01	
		P. alba	0.79 ± 0.01	0.80 ± 0.00	0.80 ± 0.01	
Parameter		Main effects			Interaction	
		Clone	Ν	Metal	$\mathbf{C} imes \mathbf{M}$	
ANOVA P valu	ies					
$V_{\rm cmax}$.0364	r	18	.0220	
$J_{\rm max}$.0242	r	18	.0035	
$J_{\rm max}/V_{\rm cmax}$		ns		0411	.0398	
lg		ns	r	18	.1772	
$\tilde{F}_{\rm v}/F_{\rm m}$		ns	r	18	ns	

Values are mean \pm standard error (n = 4). Units are expressed as μ mol of CO₂ m⁻² s⁻¹ (V_{cmax}), μ mol of electrons m⁻² s⁻¹ (J_{max}), % (l_g), relative units (F_v/F_m). ANOVA: P values for main effects and interactions clone \times metal (C \times M) are shown.



Fig. 1. Photosynthetic rate (*A*), stomatal conductance (g_s) and intercellular CO₂ concentration (C_i) in leaves of *Populus x canadensis*, *Adda* clone (black bars) and *Populus alba*, *Villafranca* clone (white bars) grown with 0.4 (control), 25 and 75 μ M Cu for 80 (T_2) days. Units are μ mol m⁻² s⁻¹ (*A*), mol m⁻² s⁻¹ (g_s), μ mol mol⁻¹ (C_i). Values are mean \pm standard error (n = 4).

thesis appeared to increase with increasing Cu treatment in *P. x canadensis*, while decreased in *P. alba*, confirming the genotypic difference of the two species in terms of gas exchange and photosynthetic capacity. Cu did not induce a significant variation in F_v/F_m values averaging 0.8 across treatments and clones (Table 5).

3.4. Copper content

Cu concentration of *P. x canadensis* and *P. alba* sampled at T_1 had similar values, averaging 8.8 and 5.9 mg kg⁻¹, respectively, in leaves and in stem of 25 and 75 μ M Cu treated plants. Cu root concentration increased in both species (Fig. 2). How-

ever, after 80 days the two species showed significant differences in term of metal allocation in leaves: in *P. alba*, Cu concentration kept increasing, while in *P. x canadensis* it averaged between 3.0 and 4.0 mg kg^{-1} through all the treatments. In both species, Cu concentration in the roots increased with the treatment.

At time T_1 , Cu content was higher in *P. x canadensis* than in *P. alba* in all plant organs (Fig. 3), and at time T_2 in stem and roots of 25 and 75 μ M Cu treated plants. Leaf Cu content at time T_2 was higher in *P. alba* than in *P. x canadensis* because of the higher roots to shoot translocation rate in this genotype.

4. Discussion

A way to evaluate the tolerance to heavy metals in plants is to measure the survival rate in heavily toxic substrates or the reduction in growth rate (Hunt, 1978) and the impairment of main physiological functions such as assimilation rate and nitrogen balance (Tognetti et al., 2004). Using this approach, in this study we observed that two different poplar species *P. x canadensis*, *Adda* clone and *P. alba*, *Villafranca* clone differ in their response to Cu treatment in term of growth and metal allocation.

P. x canadensis proved to be more tolerant to high Cu: plant growth was significantly reduced only at concentration of 75 μ M Cu, while no differences were observed between control and 25 μ M Cu treated plants. On the other hand, *P. alba* showed a remarkable reduction of growth throughout all levels of treatment.

It has been long known that in plants the application of toxic concentrations of metal can induce growth reduction because of the interference with nutrient uptake, and photosynthetic activity. Cu affects nitrogen uptake as it impairs plasma membrane functionality (Burzyński and Buczek, 1994), and it also affects nitrogen assimilation by blocking the cytosolic enzyme nitrate reductase (Kłobus et al., 2002). In accordance with these observations we found that nitrogen concentration decreased in plants supplied with Cu (Table 3), thus implying a possible interference between the metal and the nitrogen uptake.

To evaluate whether the two poplar species were similarly affected by Cu treatment in term of photosynthetic efficiency, we measured the chlorophyll content and the main photosynthetic parameters. It is known that interveinal chlorosis of leaves is one of the firstly visible symptoms of Cu toxicity (Taylor and Foy, 1985), and that it is due to decreased rate of chlorophyll biosynthesis and chlorophyll content, caused by damages to thylacoid membranes (Quartacci et al., 2000). In tolerant plants, however, it has been reported that chlorophyll content increases or does not significantly change in response to treatment with metals (Baszyński et al., 1982; Stiborova et al., 1986). When we measured the chlorophyll content after 80 days of Cu treatment, we noticed an increase in Chl_a and Chl_b content in P. x canadensis grown with 25 and 75 µM Cu, and a decrease in P. alba grown with 75 µM Cu (Table 4). That points out a higher Cu tolerance in P. x canadensis than in P. alba. Cu in leaves causes toxicity mainly because it negatively interferes with the electron transport chain and the oxygen evolving complex of the PSII (Romanowska, 2002). Moreover, it induces a decrease of initial (F_o) and maximal (F_m) fluorescence (Boucher and Carpentier, 1999), and it inhibits variable fluorescence (F_v) (Ouzounidou et al., 1993). No significant differences were reported for F_v and F_m in the two poplar species (Table 5); however, light-saturated rate of electron transport (J_{max}) and maximum rate of carboxylation (V_{cmax}) revealed that the two clones had different responses to Cu (Table 5; interaction C × M; P = .0398). In *P. x canadensis, J_{max}* and V_{cmax} slightly increased along the treatment, while the opposite was reported for *P. alba*. Accordingly, the photo-

synthetic rate increased in *P. x canadensis* while it decreased in *P. alba* (Fig. 1), possibly explaining the reduction of growth reported in *P. alba*. Altogether these results showed that the photosynthetic apparatus of *P. alba* is more sensitive to high Cu than that of *P. x canadensis*, in which the stomata conductance only seemed to be slightly affected by the treatment with the metal (Fig. 1). An explanation of these different behaviours was found by measuring Cu concentration in different plant organs, which revealed that the two species differed in term of metal allocation. Cu concentration in roots increased with increasing Cu treatment



Fig. 2. Cu concentration in leaves, stem and roots of *P. x canadensis, Adda* clone (black bars) and *P. alba, Villafranca* clone (white bars) grown with 0.4 (control), 25 and 75 μ M Cu for 30 (T_1) or 80 (T_2) days. Units are mg kg⁻¹. Values are mean \pm standard error (n=4). ANOVA: *P* values for main effects and interactions clone \times metal (C \times M), clone \times time (C \times T), metal \times time (M \times T), clone \times metal \times time (C \times M \times T) are shown.

level in both species, while in leaf it increased only in *P. alba* (Fig. 2; time T_2). Plants that preferentially accumulate heavy metals in roots generally show a wide range of metal tolerance, as the photosynthetic apparatus is protected from toxicity (Pulford et al., 2002). Accordingly photosynthesis was highly impaired in *P. alba*, while *P. x canadensis* retained good photosynthesis capacity. On the other hand, heavy metals accumulated in roots often imply root toxicity, resulting in a reduction of root growth. Root biomass reduction and changes in root morphology have

been described as a feature of Cu toxicity (Arduini et al., 1995; Patterson and Olson, 1983), while the redistribution of roots to less contaminated zones of soil has been reported as an example of facultative tolerance in trees (Pulford and Watson, 2003). In our study we observed both responses: reduction of root biomass and root thickening was observed at increasing levels of Cu in *P. alba*, while root growth seemed marginally affected in *P. x canadensis*. At the same time, at high Cu concentration, the root system in *P. x canadensis* developed preferentially on the pot



Fig. 3. Total Cu content in leaves, stem and roots of *P. x canadensis, Adda* clone (black bars) and *P. alba, Villafranca* clone (white bars) grown with 0.4 (control), 25 and 75 μ M Cu for 30 (T_1) or 80 (T_2) days. Units are mg. Values are the mean \pm standard error (n=4). ANOVA: *P* values for main effects and interactions clone \times metal (C \times M), clone \times time (C \times T), metal \times time (M \times T), clone \times metal \times time (C \times M \times T) are shown in the table.

surface instead than deeply, which may be interpreted as a plant attempt to avoid metal toxicity.

The tolerance of *P. x canadensis* towards high Cu concentrations ($TI_{25} = 93.0$ and $TI_{75} = 77.2$) was remarkable, as well as its ability to accumulate the metal in the root system without suffering from toxicity. The wide and well-expanded root apparatus of this clone could be advantageously exploited in phytostabilization programs to avoid the spreading of pollutants through the environment. Because the accumulation of Cu in stems of plants treated with 25 μ M Cu was quite high as well (0.08 mg on average; Fig. 3), the utilization of *P. x canadensis* in phytoextraction processes could also be feasible. Furthermore, as Cu concentration and total Cu content in leaves did not increase with increasing metal concentration in the treatment (Figs. 2 and 3; leaves at time T_2), the return of the metal to soil by annual leaf shedding could be avoided.

Tolerance to Cu was lower ($TI_{25} = 42.7$ and $TI_{75} = 26.3$) in P. alba, as expected in species that accumulate heavy metals in shoots more than in roots, while the root to shoot translocation rate was high, as desired for phytoextraction technologies. When heavy metals are accumulated in leaves, the risk of spreading and maintaining pollutants into the environment by leaf fall becomes an important consideration. Mertens et al. (2004) noticed that for this reason white poplar is not suitable for Cd and Zn phytoextraction. However, plants with a specific leaf accumulation profile are often used as biomonitors, because they give information on the quality of the environment (Markert et al., 2003). Biomonitoring using trees has been extensively exploited to evaluate the level of environmental pollution from soil and atmosphere, as well as the pollutant availability to living organisms. Among trees, poplars have been successfully employed in USA and in EU (Djingova et al., 1999; Wagner, 1993). Madejón et al. (2004) used P. alba leaves to monitor trace elements in soil contaminated by mine spill-out and they found that Cu concentration in leaves correlates with metal availability in surface and deep soil layers. We report here that P. alba, Villafranca clone and the hybrid P. x canadensis, Adda clone responded differently to high Cu treatments in terms of tolerance and metal allocation. P. x canadensis, Adda clone, accumulating Cu mainly in roots, present a trait sought in plants suitable for phytostabilization, while P. alba, Villafranca clone, with a high root to shoot translocation rate, could be exploited as an indicator of Cu polluted soils.

Acknowledgments

We thank the Istituto Sperimentale per la Pioppicoltura (CRA-MIPAF, Casale Monferrato, Italy) for providing cuttings of *P. x canadensis* (*Adda* clone) and *P. alba* (*Villafranca* clone). We are grateful to Matilde Paino D'urzo for critically reading the manuscript. This research was supported by grant from Italian MIUR (PRIN).

References

Álvarez, E., Fernández Marcos, M.L., Vaamonde, C., Fernández-Sanjurjo, M.J., 2003. Heavy metals in the dump of an abandoned mine in Galicia (NW) Spain) and in the spontaneously occurring vegetation. Sci. Total Environ. 313, 185–197.

- Arduini, I., Godbold, D.L., Onnis, A., 1995. Influence of copper on root growth and morphology of *Pinus pinea* L. and *Pinus pinaster* Ait. seedlings. Tree Physiol. 15, 411–415.
- Arnon, D.I., Hoagland, D.R., 1940. Crop production in artificial culture solutions and in soils with special reference to factors influencing yields and absorption of inorganic nutrients. Soil Sci. 50, 463–483.
- Baszyński, T., Król, M., Krupa, Z., Ruszkowska, M., Wojcieska, U., Wolińska, D., 1982. Photosynthetic apparatus of spinach exposed to excess of copper. Z. Pflanzenphysiol. 108, 385–392.
- Boucher, N., Carpentier, R., 1999. Hg²⁺, Cu²⁺, Pb²⁺-induced changes in photosystem II photochemical yield and energy storage in isolated thylakoid membranes: a study using simultaneous fluorescence and photoacustic measurements. Photosynth. Res. 59, 167–174.
- Burzyński, M., Buczek, J., 1994. The influence of Cd, Pb, Cu and Ni on NO₃⁻ uptake by cucumber seedlings. II. In vivo and in vitro effects of Cd, Pb, Cu and Ni on the plasmalemma ATPase and oxidoreductase from cucumber seedlings roots. Acta Physiol. Plant. 16, 297–302.
- Di Baccio, D., Tognetti, R., Sebastiani, L., Vitagliano, C., 2003. Responses of *Populus deltoides × Populus nigra (Populus × euramericana)* clone I-214 to high zinc concentrations. New Phytol. 159, 443–452.
- Dickmann, D.I., 1971. Photosynthesis and respiration by developing leaves of cottonwood (*Populus deltoides* Bartr). Bot. Gaz. 132, 252–259.
- Djingova, R., Wagner, G., Kuleff, I., 1999. Screening of heavy metal pollution in Bulgaria using *Populus nigra* "Italica". Sci. Total Environ. 234, 175–184.
- Farquhar, G.D., Sharkey, T.D., 1982. Stomatal conductance and photosynthesis. Ann. Rev. Plant Physiol. 33, 317–345.
- Farquhar, G.D., Shulze, E.D., Küppers, M., 1980. Responses to humidity by stomata of *Nicotiana glauca* L. and *Corylus avellana* L. are consistent with the optimization of carbon dioxide uptake with respect to water loss. Aust. J. Plant Physiol. 7, 315–327.
- Fischerová, Z., Tlustoš, P., Száková, J., Šichorová, K., 2006. A comparison of phytoremediation capability of selected plant species for given trace elements. Environ. Pollut. 144, 93–100.
- Harley, P.C., Sharkey, T.D., 1991. An improved model of C3 photosynthesis at high CO₂: reversed O₂ sensitivity explained by lack of glycerate reentry into the chloroplast. Photosynth. Res. 27, 169–178.
- Harley, P.C., Loreto, F., Di Marco, G., Sharkey, T.D., 1992a. Theoretical considerations when estimating the mesophyll conductance to CO₂ flux by analysis of the response of photosynthesis to CO₂. Plant Physiol. 98, 1429–1436.
- Harley, P.C., Thomas, R.B., Reynolds, J.F., Strain, B.R., 1992b. Modelling photosynthesis of cotton grown in elevated CO₂. Plant Cell Environ. 15, 271–282.
- Hrudey, S.E., Chen, W., Rousseaux, C.G., 1995. Bioavailability in Environmental Risk Assessment. CRC Lewis publishers, New York.
- Hunt, R., 1978. Plant Growth Analysis. Camelot Press Ltd., Southampton, UK.
- Jordan, D.B., Ogren, W.L., 1984. The CO₂/O₂ specificity of ribulose 1,5bisphosphate carboxylase/oxygenase. Planta 161 (4), 308–313.
- Kłobus, G., Burzyński, M., Buczek, J., 2002. Heavy metal and nitrogen metabolism. In: Prasad, M.N.V., Strzałka, K. (Eds.), Physiology and Biochemistry of Metal Toxicity and Tolerance. Kluwer Academic Publishers, The Netherlands, pp. 326–327.
- Kopponen, P., Utrianen, M., Lukkari, K., Suntioinen, S., Kärenlampi, L., Kärenlampi, S., 2001. Clonal differences in copper and zinc tolerance of birch in metal-supplemented soils. Environ. Pollut. 112, 89–97.
- Landberg, T., Greger, M., 2002. Differences in oxidative stress in heavy metal resistant and sensitive clones of *Salix viminalis*. J. Plant Physiol. 159, 69–75.
- Lasat, M.M., 2002. Phytoextraction of toxic metal: a review of biological mechanisms. J. Environ. Qual. 31, 109–120.
- Laureysens, I., De Temmerman, L., Hastir, T., Van Gysel, M., Ceulemans, R., 2005. Clonal variation in heavy metal accumulation and biomass production in a poplar coppice culture. II. Vertical distribution and phytoextraction potential. Environ. Pollut. 133, 541–551.
- Madejón, P., Marañon, T., Murillo, J.M., Robinson, B., 2004. White poplar (*Populus alba*) as a biomonitor of trace elements in contaminated riparian forests. Environ. Pollut. 132, 145–155.

- Maegher, R.B., 2000. Phytoremediation of toxic elemental and organic pollutants. Curr. Opin. Plant Biol. 3, 153–162.
- Markert, B.A., Breure, A.M., Zechmeister, H.G., 2003. Definitions, strategies and principles for bioindication/biomonitoring of the environment. In: Markert, B.A., Breure, A.M., Zechmeister, H.G. (Eds.), Bioindicators and Biomonitors. Elsevier Science, Amsterdam, pp. 3–39.
- Mertens, J., Vervaeke, P., De Schrijver, A., Luyssaert, S., 2004. Metal uptake by young trees from dredged brackish sediment: limitations and possibilities for phytoextraction and phytostabilization. Sci. Total Environ. 326, 209–215.
- Moran, R., 1982. Formulae for determination of chlorophyllous pigments extracted with *N*,*N*-dimethylformamide. Plant Physiol. 69, 1376–1381.
- Ouzounidou, G., Lannoye, R., Karataglis, S., 1993. Photoacoustic measurements of photosynthetic activities in intact leaves under copper stress. Plant Sci. 89, 221–226.
- Patterson, W., Olson, J.J., 1983. Effects of heavy metals on radicle growth of selected woody species germinated on filter paper, mineral and organic soil substrates. Can. J. Forest Res. 13, 233–238.
- Pulford, I.D., Watson, C., 2003. Phytoremediation of heavy metal-contaminated land by trees — a review. Environ. Int. 29, 529–540.
- Pulford, I.D., Riddell-Black, D., Stewart, C., 2002. Heavy metal uptake by willow clones from sewage sludge-treated soil: the potential for phytoremediation. Int. J. Phytoremediation 4, 59–72.
- Punshon, T., Lepp, N.W., Dickinson, N.M., 1995. Resistance to copper toxicity in some British willows. J. Geochem. Explor. 52, 259–266.
- Quartacci, M.F., Pinzino, C., Sgherri, C.L.M., Navari-Izzo, F., 2000. Growth in excess copper induces changes in the lipid composition and fluidity of PSII-enriched membranes in wheat. Physiol. Plantarum 108, 87–93.
- Rokwood, D.L., Naidu, C.V., Carter, D.R., Rahmani, M., Spriggs, T.A., Lin, C., Alker, G.R., Isebrands, J.G., Segrest, S.A., 2004. Short-rotation woody crops

and phytoremediation: opportunities for agroforestry? Agroforest. Syst. 61, 51–63.

- Romanowska, E., 2002. Gas exchange functions in heavy metal stressed plants. In: Prasad, M.N.V., Strzałka, K. (Eds.), Physiology and Biochemistry of Metal Toxicity and Tolerance. Kluwer Academic Publishers, The Netherlands, pp. 260–262.
- Salt, D.E., Smith, R.D., Raskin, D., 1998. Phytoremediation. Ann. Rev. Plant Physiol. Plant Mol. Biol. 49, 643–668.
- Sebastiani, L., Scebba, F., Tognetti, R., 2004. Heavy metal accumulation and growth responses in poplar clones Eridano (*Populus deltoids × maximowiczii*) and I-214 (*P. × euramericana*) exposed to industrial waste. Environ. Exp. Bot. 52, 79–88.
- Sharkey, T.D., 1985. Photosynthesis in intact leaves of C₃ plants: physics, physiology and rate limitations. Bot. Rev. 51, 53–105.
- Stiborova, M., Dubravova, M., Brezinova, A., Friedrich, A., 1986. Effect of heavy metal ions on growth and biochemical characteristics of photosynthesis of barley. Photosynthetica 20, 418–425.
- Taylor, G.J., Foy, C.D., 1985. Differential uptake and toxicity of ionic and chelated copper in *Triticum aestivum*. Can. J. Bot. 63, 1271–1275.
- Tognetti, R., Sebastiani, L., Minnocci, A., 2004. Gas exchange and foliage characteristics of two poplar clones grown in soil amended with industrial waste. Tree Physiol. 24, 75–82.
- von Caemmerer, S., Farquhar, G.D., 1981. Some relationship between the biochemistry of photosynthesis and the gas exchange of leaves. Planta 153, 376–387.
- Wagner, G.J., 1993. Accumulation of cadmium in crop plants and its consequences to human health. Adv. Agron. 51, 173–212.
- Wullschleger, S.D., 1993. Biochemical limitations to carbon assimilation in C₃ plants—a retrospective analysis of the A/C_i curves from 109 species. J. Exp. Bot. 44, 907–920.