

See discussions, stats, and author profiles for this publication at: <https://www.researchgate.net/publication/260133796>

On the Empirical Content of Carbon Leakage Criteria in the EU Emissions Trading Scheme.

Article in *Ecological Economics* · December 2013

Impact Factor: 2.72 · DOI: 10.1016/j.ecolecon.2014.05.010

CITATIONS

8

READS

67

4 authors:



[Ralf Martin](#)

Imperial College London

50 PUBLICATIONS 588 CITATIONS

[SEE PROFILE](#)



[Mirabelle Muûls](#)

Imperial College London

24 PUBLICATIONS 269 CITATIONS

[SEE PROFILE](#)



[Laure de Preux](#)

Imperial College London

11 PUBLICATIONS 98 CITATIONS

[SEE PROFILE](#)



[Ulrich j Wagner](#)

University Carlos III de Madrid

26 PUBLICATIONS 236 CITATIONS

[SEE PROFILE](#)

ON THE EMPIRICAL CONTENT OF CARBON LEAKAGE CRITERIA IN THE EU EMISSIONS TRADING SCHEME*

Ralf Martin[†]
Mirabelle Muûls[‡]
Laure B. de Preux[§]
Ulrich J. Wagner^{**}

December 2013

Abstract

The EU Emissions Trading Scheme continues to exempt industries deemed at risk of carbon leakage from permit auctions. Carbon leakage risk is established based on the carbon intensity and trade exposure of each 4-digit industry. Using a novel measure of carbon leakage risk obtained in interviews with almost 400 managers at regulated firms in six countries, we show that carbon intensity is strongly correlated with leakage risk whereas trade exposure is not. In spite of this, most exemptions from auctioning are granted to trade exposed industries. Our analysis suggests two ways of tightening the exemption criteria without increasing relocation risk among non-exempt industries. The first one is to exempt trade exposed industries only if they are also carbon intensive. The second one is to consider exposure to trade only with less developed countries. By modifying the carbon leakage criteria along these lines, European governments could raise additional revenue from permit auctions of up to €3 billion per year.

Keywords: Carbon leakage, industrial relocation, emissions trading, EU ETS, permit allocation, firm data

JEL Classifications: H23, H25, Q52, Q54, F18

* Part of the analysis in this paper was previously circulated as NBER Working Paper 19097 (Martin et al., 2013). We are indebted to Barry Anderson, Jörg Leib and Marty McGuigan for their invaluable help at different stages of this research, and to Felix de Bousies, Pieter De Vlieger, David Disch, Eszter Domokos, Lorenz Elsasser, Helen Franzen, Maite Kervyn, Zsofia Kopetka, Oliwia Kurtyka, Anne-Lise Laurain, Emeric Lujan, Nicole Polsterer, Antoine Martin-Regniault, Maxence Snoy, Joanna Romanowicz, Bartosz Vu, Julia Wittig, Joanna Wylegala for their help with the interviews. Melanie Hermann and Antonin Cura provided excellent research assistance. Olivier Sartor, Stephen Lecourt and Clément Pallière graciously provided data on National Implementation Measures. We have received helpful comments from Stephen Boucher, Alex Bowen, Denny Ellerman, Sam Fankhauser, Tom Foxon, Andy Gouldson, Mark Jacobsen, Stéphanie Monjon, and from staff members at DECC, at DG Climate, and at the Environmental Committee of the European Parliament. All remaining errors are our own. The interviews were funded through grants from the European Climate Foundation and the ESRC. The Centre for Economic Performance and the Grantham Institute on Climate Change provided generous logistical support. The authors gratefully acknowledge financial support from the British Academy (Martin), from the Leverhulme Trust (Muûls) and from the Spanish Ministry for Science and Innovation, reference number SEJ2007-62908 (Wagner).

[†] Imperial College Business School, South Kensington Campus, London SW7 2AZ, United Kingdom, Grantham Institute on Climate Change, and Centre for Economic Performance (CEP), London School of Economics (LSE).

Email: r.martin@imperial.ac.uk

[‡] Grantham Institute for Climate Change and Imperial College Business School, South Kensington Campus, London SW7 2AZ, United Kingdom, and CEP. Email: m.muuls@imperial.ac.uk

[§] Imperial College Business School, South Kensington Campus, London SW7 2AZ, United Kingdom, and CEP, LSE.

Email: l.depreux@imperial.ac.uk

^{**} Departamento de Economía, Universidad Carlos III de Madrid, Calle de Madrid 126, 28903 Madrid, Spain. Email: uwagner@eco.uc3m.es

1 Introduction

It is widely recognized that the problem of carbon leakage poses a major challenge for designing effective unilateral policies aimed at mitigating global climate change. In its most direct manifestation, carbon leakage occurs when polluting plants that are subject to climate policy relocate to an unregulated jurisdiction. Since carbon emissions are a global pollutant, their “leaking” to unregulated places reduces the environmental benefits from the policy. In addition, carbon leakage creates an excess burden for those countries that regulate emissions to the extent that relocation reduces output, employment, and taxable profits at home.

Not surprisingly, carbon leakage takes the center stage whenever new climate change regulation is up for debate. So far, the most common deterrent against carbon leakage has been to either compensate or to exempt those industries deemed to be most adversely affected by the policy. For instance, virtually all of the numerous carbon taxes that have emerged in Europe since the 1990’s grant rebates or exemptions to energy-intensive firms in order to prevent them from relocating.¹ While this practice can be justified from the point-of-view of industrial policy, it runs counter to the polluter-pays principle underlying environmental policy-making in the EU. It also gives way to rent-seeking behavior, as regulated firms have an incentive to exaggerate their compliance costs in order to receive more generous compensation. Addressing carbon leakage is therefore a difficult and controversial policy issue.

In this paper, we analyze the current scheme to prevent carbon leakage implemented in the European Emissions Trading System (EU ETS), the world’s first and largest regional cap-and-trade system for greenhouse gas emissions. During the first eight years of the EU ETS, leakage was addressed by offering manufacturing firms generous compensation in the form of allocating most emission allowances free of charge. In the current, third trading phase, which runs from 2013 until 2020, the European Commission (EC) gradually reduces the proportion of free allowances allocated to manufacturing firms. At the same time, and contrary to its stated objective of achieving full auctioning of emission allowances, the EC exempts from this transition more than three quarters of the regulated emissions from manufacturing, on the grounds that

¹In their analysis of the UK Climate Change Levy, Martin, de Preux, and Wagner (2011) find no causal impact on output, employment or plant exit among treated firms.

the firms accounting for those emissions are at risk of carbon leakage. Exemptions are granted according to two simple criteria, namely the carbon intensity of value added and trade exposure, measured at the level of the 4-digit industry code.

Our paper assesses the accuracy of these criteria based on a novel firm-level measure of leakage risk we gathered in telephone interviews with managers of 761 manufacturing firms in six European countries ([Martin, Muûls, de Preux, and Wagner, 2014](#)). The flexibility of the interview based approach, along with the bias-reducing format of the survey tool developed by Bloom and van Reenen (2007) and adapted to the climate policy context in [Martin et al. \(2012, 2014\)](#), allows us to elicit valuable information on politically contentious issues such as a firm's vulnerability to carbon pricing, defined as the firm's propensity to downsize or relocate in response to climate change policy.

We show that carbon intensity is strongly correlated with our interview-based measure of vulnerability whereas trade intensity is not. This is a reason for concern because most exemptions from auctioning are granted on the basis of the trade intensity criterion alone. We propose two simple improvements to the exemption criteria, based on the principle that free permits should only be given to industries where the average relocation propensity is significantly higher than that of non-exempt industries. First, by not exempting trade intensive sectors but the ones that are at least moderately carbon intensive as well, European governments could raise additional auction revenue of up to €3 billion every year. Second, we show that a sector's intensity of trade with less developed countries such as China is a better proxy for vulnerability than the overall trade intensity. A change in the definition of the trade intensity criterion along these lines could raise an additional €430 million in auction revenues per year.

In extending the normative analysis of industry compensation rules in the EU ETS by [Martin et al. \(2014\)](#), this paper contributes further evidence of practical value on this controversial aspect of climate policy. This will be relevant for the revision of the carbon leakage criteria by the EU Commission in 2014, but our findings also inform climate policy far beyond the European context. This is because similar criteria to the ones used by the EC have been adopted in actual and proposed legislation underlying half a dozen regional carbon trading schemes world wide. For instance, emission intensity and trade intensity are used to determine eligi-

bility for compensation in the recently implemented carbon trading schemes in California and Switzerland, in Australia's Carbon Pollution Reduction Scheme and in New Zealand's ETS. Moreover, these metrics were proposed for a US wide cap-and-trade scheme under the 2009 Waxman-Markey Bill, and will be applied in a future South Korean ETS (cf. [Hood, 2010](#); www.icapcarbonaction.com). In view of this, it seems of first-order importance to know how these criteria relate to leakage risk by the very managers who get to decide on relocation.

The next section describes the policy background and summarizes the relevant literature. Section 3 describes the data set and explains our regression based test. Section 4 presents the results and Section 5 discusses their implications for the auction revenues forgone by the actual policy. Section 6 concludes.

2 Policy background: Carbon Leakage and the EU ETS

2.1 Permit allocation, benchmarking and carbon leakage sectors

According to the independence property of emissions trading ([Montgomery, 1972](#)), the permit price only depends on the stringency of the overall cap, but not on the initial allocation. An implication of this is that the permit allocation *per se* does not condition firm behavior at the intensive margin, because firms factor the opportunity cost of using a permit into their marginal cost – regardless of the initial cost of acquiring permits. Independence need not hold in the presence of market power ([Hahn, 1984](#)) or transaction costs ([Stavins, 1995](#)), but existing research has not rejected the independence property in the EU ETS ([Convery and Redmond, 2007](#); [Reguant and Ellerman, 2008](#)).

In contrast, the extensive-margin behavior of firms is affected by both the overall stringency and the initial permit allocation. This is because variable profits are decreasing in the permit price and total profits decrease with the total cost of permits. Thus, full auctioning of emission permits might lower firm profits to the point where exit or relocation to non-EU countries are worth considering. Likewise, a more stringent cap may have similar effects as it sustains a higher carbon price. An economic rationale behind allocating permits for free is thus to mitigate the risk of relocation and carbon leakage by compensating industry for the adverse profit

impacts of emissions trading.

Initial permit allocation in phases I and II of the EU ETS was delegated to the member states who drew up a National Allocation Plan (NAP) that both fixed the national cap and determined the sectoral allocation. The principles guiding NAP development in phase I were quite consistent across countries, as most opted for “grandfathering”, i.e. free permit allocations based on historical emissions ([Ellerman, Buchner, and Carraro, 2007](#)). In phase II, the member states imposed more stringent caps so as to honor their commitment to the EU’s joint emission target under the Kyoto Protocol, but they also retained free allocation. Auctioning fell far short of what was allowed, and benchmarking remained an exception ([Ellerman and Joskow, 2008](#)).

In phase III, the allocation of free emission allowances is relegated from national governments back to Brussels. Directive 2009/29/EC advances the transition towards auctioning of permits as the basic principle of allocation and stipulates a harmonized allocation scheme to reduce competitive distortions among producers of similar products across member states. The two main features of this scheme are (i) the use of benchmarks which rewards operators who have taken early action to reduce the emission intensity of production, and (ii) the continued free allocation to sectors considered at risk of carbon leakage.

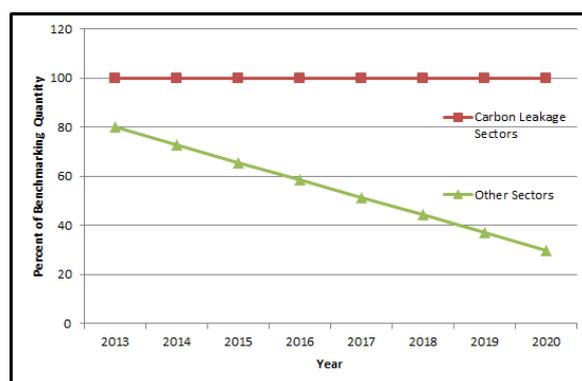
To the extent possible, benchmarks are defined in tons of CO₂ equivalent per unit of output of a specific product.² They reflect the average greenhouse gas emission performance of the 10% best performing installations in the EU producing that product, based on the average emissions intensity in 2007-2008.³ The amount of free permits is obtained by multiplying the benchmark with the historical reference activity level, defined as the median activity level over the years from 2005 until 2008 (or from 2009 until 2010, if larger). Total allocations calculated in this way are scaled by a factor that takes a value of 0.8 in 2013 and declines linearly to a factor of 0.3 in 2020.⁴ This factor is meant to accomplish the gradual transition to full auctioning foreseen already in the first version of the EU Emissions Trading Directive.

²Where deriving a product benchmark is not feasible, a hierarchy of fallback approaches is applied, as explained in detail by Lecourt, Pallière, and Sartor (2013).

³Cf. Commission Decision 2011/87/EU determining transitional Union-wide rules for harmonized free allocation of emission allowances pursuant to Article 10a of Directive 2003/87/EC of the European Parliament and of the Council (2011) OJ L 130/1 (Benchmarking Decision).

⁴Furthermore, a uniform correction factor is applied if necessary to align the total free allocation to benchmarked installations with the overall cap on emissions.

Figure 1: Free Allocation under Benchmarking

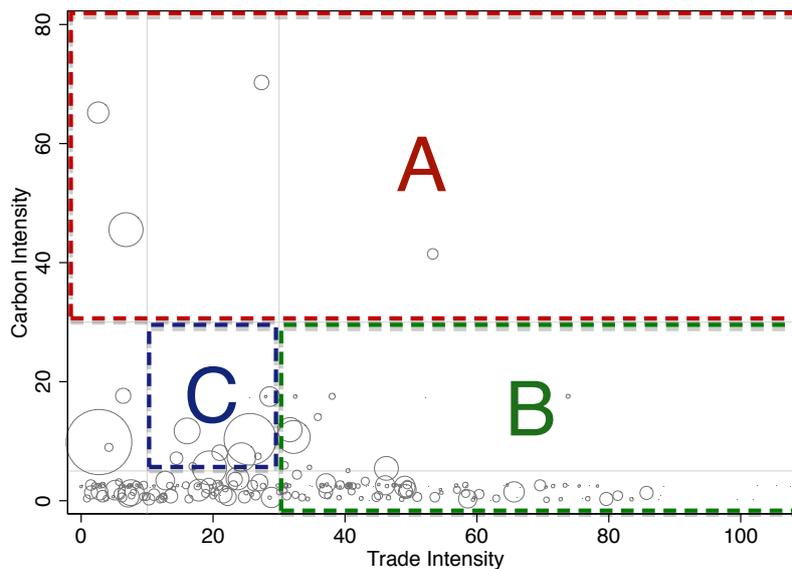


However, carbon intensive industries convinced EU law makers that full auctioning of permits would have a detrimental impact on their competitiveness. In response to this, the EC grants 100% of benchmark allocations for free to firms in sectors that are considered at risk of carbon leakage, as depicted in Figure 1. The Carbon Leakage Decision⁵ stipulates that leakage risk of a sector or subsector be assessed on the basis of its carbon intensity (CI) and/or trade intensity (TI). The former proxies for the cost burden imposed by full auctioning, and is measured as the sum of the direct and indirect costs of permit auctioning, divided by the gross value added of a sector. The direct costs are calculated as the value of direct CO₂ emissions, where a proxy price of 30€/tCO₂ is used. Indirect costs measure the exposure to electricity price increases that are inevitable on account of full permit auctioning in the power sector. To calculate indirect costs, electricity consumption (in MWh) is multiplied by the average emission intensity of electricity generation in the EU27 countries (0.465 tCO₂/MWh), and by the same proxy price of 30€/tCO₂ for one European Union Allowance. The TI metric is calculated as “the ratio between the total value of exports to third countries plus the value of imports from third countries and the total market size for the Community (annual turnover plus total imports from third countries)” (EU Commission, 2009, p. 24).

According to Directive 2009/29/EC, a combination of thresholds for CI and TI is used to assess carbon leakage risk. For a sector to be considered at significant risk of carbon leakage, its CI must be greater than 5% and its TI greater than 10%, or else its CI or TI is greater than 30%. This enables us to classify eligible sectors into three mutually exclusive categories:

⁵Commission Decision 2010/2/EU determining, pursuant to Directive 2003/87/EC of the European Parliament and of the Council, a list of sectors and subsectors which are deemed to be exposed to a significant risk of carbon leakage (2010) OJ L 1/10 (Carbon Leakage Decision).

Figure 2: Sectors exempt from permit auctions



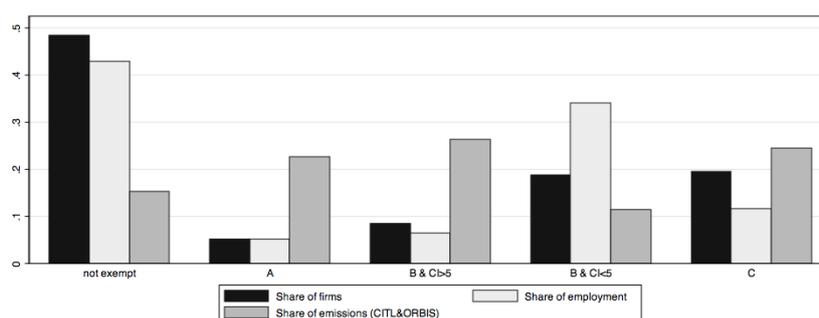
Notes: The figure shows a scatter plot of the carbon and trade intensities of 4-digit (NACE 1.1) manufacturing industries, based on 9,061 EU ETS installations. The size of the circles is proportional to the number of firms in a given industry. Sectors in areas A, B, and C will continue to be exempt from permit auctions in EU ETS phase III. Source: Martin et al. (2014).

- **A:** high carbon intensity ($CI > 30\%$)
- **B:** high trade intensity and low to moderate carbon intensity ($CI \leq 30\% \cap TI > 30\%$),
- **C:** moderate carbon and trade intensities ($5\% < CI \leq 30\% \cap 5\% < TI \leq 30\%$).

Figure 2 plots the location of 3-digit sectors in a diagram with CI on the vertical and TI on the horizontal axis.⁶ Two findings strongly emerge from this graph: First, category B contains most of the sectors the EC considers at risk of carbon leakage. Second, most of these carbon leakage sectors are not carbon intensive at all, as their carbon intensity is less than 5%. After splitting category B according to its carbon intensity we plot in Figure 3 the relative size of the resulting five categories in terms of the shares in the number of firms, in employment and in CO₂ emissions. By all these measures, category B turns out to be the largest group of exempted firms. The share of CO₂ emissions that is not exempt from auctioning is as small as 15%, which is in line with an alternative estimate of 23% by [Juergens et al. \(2013\)](#). This means that the Carbon Leakage Decision leaves most pollution rights with European industry and hence strongly undermines the principle of full auctioning established in the amended ETS directive.

⁶Similar graphical representations have been used by Clò (2010) and Martin et al. (2014).

Figure 3: Relative size of the exemption groups



Notes: The chart displays the relative size of each group of NACE industries which are defined by the exemption criteria. Category B (very trade intensive sectors) is subdivided into low and moderate carbon intensity. The sample includes the 4,254 manufacturing firms participating in the EU ETS and matched to ORBIS. The first bar indicates a group's share in the total number of firms, the second bar its share in employment, and the third bar its share in CO₂ emissions, based on the number of surrendered permits recorded in the CITL. To compute CI and TI figures at the NACE 4-digit level, we follow the methodology and databases used by the EU Commission (2009). Source: Martin et al. (2014).

2.2 How do the carbon leakage criteria relate to profit impacts?

Conceptually, one can identify as relevant aspects (i) the cost impact of regulation, (ii) the demand response to higher product prices, and (iii) the factor specificity of production. The cost impact stems from the fact that previously grandfathered firms will be forced to pay the market price for the right to pollute. The cost burden tends to be higher for firms with a higher ratio of direct and indirect emissions to gross value added. However, CI defined in this way is an incomplete measure of the cost impact, because it fails to account for how easily carbon intensive inputs can be replaced by less carbon intensive ones.

The demand response determines a firm's ability to pass on the cost impact to its consumers. Charging higher prices is more difficult when customers can easily substitute to relatively cheaper products from competitors located outside the EU. The import component of the TI metric picks up this kind of demand response; in fact import penetration is a widely used proxy for firms' ability to pass-through cost increases to customers. But the TI metric also contains the export ratio, whose link to the demand response is ambiguous. On the one hand, a sector with a high export share is hit harder when losing market share to competitors in non-EU markets as a result of carbon pricing. On the other hand, a higher export intensity also signals that production in the EU benefits from specific factors not easily available in other locations, and this tends to mitigate the profit impact of permit auctioning. To the extent that proximity to natural resource deposits, labor pooling, or externalities from industrial agglomeration are important for certain industries, firms in these sectors are unwilling to relocate and will prefer to

serve foreign markets through exporting, causing TI to be high. Consequently, the sector looks vulnerable according to EU criteria even though it can easily pass through the cost of permit auctioning in international product markets.

Another measurement issue arises from the aggregation to the 4-digit sector level, which lumps together many different products. [Clò \(2010\)](#) points out that, in spite of a high import penetration at the sector level, firms may be able to pass through the costs of carbon in some product markets that are less competitive due to concentration or product differentiation.

Theoretical research suggests that the EU ETS adversely impacts on production in most regulated industries while rising electricity prices lower the profitability of highly exposed industries such as primary aluminum production ([Demailly and Quirion, 2006, 2008](#); [Reinaud, 2005](#)).⁷ These studies also show that free permit allocation offsets negative profit impacts in most industries, and can even lead to overcompensation ([Smale, Hartley, Hepburn, Ward, and Grubb, 2006](#)). Based on this literature, [Sato, Grubb, Cust, Chan, Korppoo, and Ceppi \(2007\)](#) propose to use trade intensity, carbon intensity and electricity intensity as proxies for the competitiveness impact of the EU ETS.

There is, however, little empirical evidence to date that directly links CI and TI to sectoral heterogeneity in the relocation response to carbon pricing. In fact, the literature on the *ex-post* analysis of the impact of the EU ETS on international competitiveness is still in its infancy. This literature, which is reviewed in detail by [Martin, Muûls, and Wagner \(2013b\)](#), does not suggest that industrial firms on the whole suffered strong adverse impacts when permits were allocated for free in the first years of the EU ETS ([Anger and Oberndorfer, 2008](#); [Abrell et al., 2011](#); [Chan et al., 2013](#); [Commins et al., 2011](#); [Bushnell et al., 2013](#); [Wagner et al., 2013a,b](#)).

Few papers so far have analyzed the link between carbon leakage and free allowance allocation in the EU ETS specifically. [Monjon and Quirion \(2011\)](#) use a computable partial equilibrium model to compare various configurations of border-tax adjustments and output-based allocation to counter carbon leakage in the EU ETS. In a theoretical analysis, [Meunier, Ponsard,](#)

⁷A widespread approach to assessing aggregate leakage effects has been to calibrate computable general equilibrium models that are capable of predicting the consequences of differential carbon pricing across regions. We do not review these models here as they are not informative about individual industries. Models with exogenous technical change predict carbon leakage rates between 5 and 35% for the Kyoto Protocol commitments ([Paltsev, 2001](#)).

and Quirion (2012) show that a combination of output based and capacity based allowance allocation is second-best when border-tax adjustments are not available. [Martin et al. \(2014\)](#) also take a normative approach, showing how free allowances should be distributed across firms in order to minimize leakage, subject to a maximum amount of foregone auction revenue. While they use microdata on free allocation in the EU ETS to conduct numerical simulations of the efficiency gains of their proposed allocation rule, this paper uses regression analysis to analyze the accuracy of the specific carbon leakage criteria used by the EC.

3 Data and Methods

3.1 Data

We use a unique firm-level data set constructed by [Martin et al. \(2014\)](#) for analyzing the link between permit allocation and carbon leakage. A key ingredient of this data set is a measure of a firm's propensity to downsize or relocate in response to carbon pricing, obtained from interviews with managers of 761 manufacturing firms in six European countries: Belgium, France, Germany, Hungary, Poland and the UK. As in [Bloom and van Reenen \(2007\)](#), interviews were conducted over the telephone and follow a protocol intended to minimize cognitive bias. Further, a large sample size and interviewer rotation is exploited to control for possible bias on the part of the interviewers by including interviewer fixed effects in regression analyses (see also [Bloom and van Reenen, 2010](#)). Specifically, managers were asked: "Do you expect that government efforts to put a price on carbon emissions will force you to outsource part of the production of this business site in the foreseeable future, or to close down completely?", and the interviewers ranked the answers to obtain an ordinal 'vulnerability score' (VS). The highest score of 5 was assigned if the manager expected the plant to be closed completely, whereas the lowest score of 1 was given if the manager expected no detrimental impacts at all. A score of 3 was assigned if the manager expected that at least 10% of production and/or employment would be outsourced in response to future policies. Scores of 2 or 4 were given to account for intermediate responses. Table 1 presents the empirical distribution of the vulnerability score (VS) across countries and industries.

Table 1: Descriptive statistics of the vulnerability score

	Mean	Standard deviation	Min	P25	Median	P75	Max	Firms
Overall vulnerability score	1.87	1.29	1	1	1	3	5	725
<i>A. by country</i>								
Belgium	1.69	1.13	1	1	1	3	5	122
France	2.07	1.34	1	1	1	3	5	136
Germany	2.12	1.58	1	1	1	3	5	131
Hungary	1.50	0.95	1	1	1	2	4	68
Poland	2.03	1.40	1	1	1	3	5	74
UK	1.75	1.12	1	1	1	3	5	194
<i>B. by 3-digit sector</i>								
Cement	2.33	1.52	1	1	1	4	5	63
Ceramics	2.15	1.46	1	1	1	3	5	13
Chemical & Plastic	1.86	1.26	1	1	1	3	5	118
Construction	1.00	0.00	1	1	1	1	1	3
Fabricated Metals	1.67	0.93	1	1	1	3	4	45
Food & Tobacco	1.56	1.01	1	1	1	2	5	106
Fuels	2.71	1.59	1	1	3	4	5	14
Furniture & NEC	1.47	0.87	1	1	1	2	4	17
Glass	2.76	1.57	1	1	3	4	5	29
Iron & Steel	2.69	1.56	1	1	3	4	5	39
Machinery & Optics	1.26	0.68	1	1	1	1	4	68
Other Basic Metals	1.78	1.39	1	1	1	2	5	9
Other Business Services	2.67	0.58	2	2	3	3	3	3
Other Minerals	3.38	1.69	1	2	4	5	5	8
Publishing	1.58	1.02	1	1	1	2	4	19
TV Communication	1.91	1.45	1	1	1	3	5	11
Textile & Leather	1.90	1.33	1	1	1	3	5	20
Vehicles	1.62	0.99	1	1	1	2	4	47
Wholesale	1.40	0.89	1	1	1	1	3	5
Wood & Paper	1.85	1.36	1	1	1	3	5	88

Notes: Summary statistics of the overall vulnerability score (first row), by country (panel A) and by 3-digit NACE sector (panel B). The score ranges from 1 (no impact) to 5 (complete relocation). A score of 3 is given if at least 10% of production of employment would be outsourced in response to future carbon pricing. NEC: Not elsewhere classified.

As [Martin et al. \(2014\)](#) point out, none of the principal manufacturing industries in the sample exhibit a significant risk of firm relocation or closure. Only Other Minerals has an average score slightly above 3, implying downsizing by at least 10% of employment or output. For a few sectors, including Iron and Steel, Ceramics, Glass, and Fuels, the 95% confidence interval includes a score of 3. In no case does the confidence band include the maximum score, meaning that the possibility of complete relocation in response to carbon pricing is very unlikely.

The interview data are augmented with “hard” data on employment and turnover from the ORBIS database, which also provides information on 3-digit NACE codes. Data from EURO-

STAT were used to reproduce as closely as possible the EC’s calculation of the sector-level variables CI and TI. Finally, firm-level data on CO₂ emissions and permit allocations was obtained from the European Union Transaction Log (EUTL, formerly known as CITL) and from the National Implementation Measures (NIM) which contain permit allocation for the current trading phase.⁸ For a comprehensive description of the data set, the interested reader is referred to [Martin et al. \(2014\)](#), which also contains a thorough analysis of the ability of the VS to capture firms’ downsizing risk.

3.2 A regression based test

To evaluate the accuracy of the EC’s carbon leakage criteria, we examine how they correlate with VS. In particular, CI and TI should be positively correlated with VS. We test this hypothesis by estimating partial correlations in a regression framework that controls for possible confounders at the firm and sector levels. The basic regression equation is given by

$$VS_{i,s} = \beta_0 + \beta_T TI_s + \beta_C CI_s + \mathbf{x}'_{i,s} \beta_{\mathbf{x}} + \varepsilon_{i,s} \quad (1)$$

where $VS_{i,s}$ is the vulnerability score of firm i in sector s , TI_s and CI_s are the EC’s trade and carbon criteria at the sector level, and $\mathbf{x}_{i,s}$ is a vector including higher order terms of these variables, country dummies, and interviewer fixed effects to control for possible bias on the part of the interviewers. Moreover, we control for interview noise due to the manager’s characteristics – by including the tenure in the company, dummies for gender and professional background (technical or law) – and due to the time of the interview – by including dummies for month, day of week and time of day (am/pm).

It could be argued that the continuous relationship between VS, CI and TI imposed in these regressions is not appropriate for the EC’s threshold based approach. We thus modify equation (1) to include a set of dummy variables representing the exemption categories (**A**, **B**, **C**) defined

⁸We thank Oliver Sartor, Stephen Lecourt and Clément Pallière for kindly providing us with the data for 20 of these countries, for which they collected and matched the NIM data on free permit allocation to ORBIS (see [Lecourt et al., 2013](#)). We complemented this dataset with the NIM data for Belgium and Hungary, which we matched to ORBIS by hand. In total, this results in a sample of nearly 8,000 installations covering 95% of the emissions.

above instead of the continuous variables TI and CI .

$$VS_{i,s} = \gamma_0 + \gamma_A \mathbb{I}_{\{i \in \mathbf{A}\}} + \gamma_B \mathbb{I}_{\{i \in \mathbf{B}\}} + \gamma_C \mathbb{I}_{\{i \in \mathbf{C}\}} + \mathbf{x}'_{i,s} \boldsymbol{\gamma}_x + \eta_{i,s} \quad (2)$$

The omitted category in this regression are firms that are not exempt under the Carbon Leakage Decision. We estimate these regressions using ordinary least squares and calculate robust standard errors which are clustered at the 4-digit NACE code level.

4 Results

4.1 Baseline specifications

Table 2 summarizes the results of various versions of regression equation (1). In the univariate specifications, we find a strong positive association of VS with carbon intensity, but no statistically significant association with trade intensity. This result is robust when both measures are included in a quadratic form that is better suited to capture possible effects of interactions and non-linearities. For instance, trade exposure could matter for very high values of TI only, or only when it coincides with high CI. There is no evidence of such effects. Weighting the regression equation (1) by employment does not change the qualitative findings but gives rise to a larger estimate for the impact on CI. This suggests that CI is a particularly good measure of the risk of downsizing among large firms. In sum, our regression-based test reveals that TI is not a good indicator to measure the risk of downsizing or outsourcing whereas CI is.

We obtain similar results when looking at exemption categories. The first column of Table 3 reports the results obtained for equation (2). Only the very carbon intensive group (**A**) has an average VS significantly higher than the reference category (firms that are not exempt from auctioning). But even in group **A** there is no dramatically high risk of downsizing or outsourcing for the average firm. The 95%-confidence band for the VS in group **A** just about includes the value of 3, which means a reduction of at least 10% in production or employment due to outsourcing.

Taken together, the regression results obtained in equations (1) and (2) suggest that the

Table 2: Vulnerability score and exemption criteria

	(1)	(2)	(3)	(4)	(5)
	Vulnerability Score (VS)				
Sectoral Trade Intensity (TI)	-0.012 (0.092)		0.050 (0.112)	0.051 (0.096)	0.097 (0.117)
Carbon Intensity (CI)		0.229*** (0.063)	0.454** (0.215)	0.292*** (0.090)	0.473*** (0.114)
TI X TI			-0.037 (0.037)		
CI X CI			0.007 (0.074)		
TI X CI			0.059 (0.106)	0.086 (0.091)	0.063 (0.134)
Weights	no	no	no	no	employment
Observations	392	392	392	392	392

Notes: OLS regressions in columns 1 to 4 and Weighted Least Squares (WLS) regression in column 5. The dataset is a cross-section of 392 interviewed firms that are part of the EU ETS and for which CITL, sectoral trade and carbon intensity data are available. The dependent variable is the vulnerability score of the firm given by the interviews data. In column 5, the score is weighted by the firm's employment. As explanatory variables, CI indicates carbon intensity and TI trade intensity which are calculated using data from Eurostat and the EU Commission. X indicates that two variables are interacted. All regressions include a constant, interview noise controls and country dummies (not reported). Robust standard errors, clustered by 4-digit NACE sector, are given in parentheses. Asterisks indicate statistical significance at the 10%(*), 5%(**) and 1%(***) level.

efficiency of the allocation scheme could be enhanced if the exemption criteria or associated thresholds were modified so as to better reflect the true risk of carbon leakage. The next section considers two simple modifications along these lines.

4.2 Extensions

4.2.1 Modifying intensity thresholds

The result that the average VS in categories **B** and **C** is not significantly higher than in sectors not exempt from auctioning suggests that subjecting sectors in these categories to auctioning would not raise overall relocation risk. However, category **B** is very heterogeneous. While most sectors in this category are not carbon intensive at all ($CI < 5$), there is a small number of sectors with intermediate carbon intensity ($5 < CI < 30$), as shown in Figure 2. In order to account for this heterogeneity, we subdivide category **B** into a group with low CI ($\mathbf{B} \cap CI < 5$) and one with intermediate CI ($\mathbf{B} \cap CI > 5$).

When these separate groups are included along with groups **A** and **C** in regression equation (1), the more carbon-intensive sectors in group **B** exhibit a significantly higher risk of outsourcing than the reference group, even though, as is the case for group **A**, the risk of downsizing or

Table 3: Vulnerability score and exemption categories

	(1)	(2)	(3)	(4)	(5)
	Vulnerability Score			Vulnerability Score>2	
CI>30 (A)	1.032*** (0.303)	1.015*** (0.312)	1.996*** (0.523)	0.714*** (0.242)	1.704*** (0.448)
TI>30 \cap CI<30 (B)	0.225 (0.258)				
10<TI<30 \cap 5<CI<30 (C)	0.122 (0.248)	0.139 (0.240)	0.358 (0.241)	0.105 (0.233)	0.271 (0.292)
B \cap CI>5		0.596* (0.316)	1.031*** (0.322)	0.500** (0.252)	1.267*** (0.417)
B \cap CI<5		-0.053 (0.243)	0.056 (0.329)	-0.059 (0.233)	0.121 (0.389)
Constant	1.623*** (0.516)	1.572*** (0.523)	1.426 (0.912)		
Weights	no	no	employment	no	employment
Observations	392	392	392	392	392

Notes: OLS regressions in columns 1 and 2, WLS in column 3 and Probit regressions in columns 4 and 5. The dataset is a cross-section of 392 interviewed firms that are part of the EU ETS and for which CITL, sectoral trade and carbon intensity data are available. The dependent variable is the vulnerability score (on a scale of 5) of the firm given by the interviews data in regressions 1 to 3, and a dummy indicating whether the score is higher than 2 in regressions 4 and 5. In columns 3 and 5, the firm's employment is used to weight the regression. CI indicates carbon intensity and TI trade intensity, calculated using data from Eurostat and the EU Commission. Based on these, dummies are constructed to represent belonging to categories A, B and C, as well as $(B \cap CI_5)$ and $(B \cap CI_{\neq 5})$. These are used as explanatory variables. Columns 4 and 5 report marginal effects of the probit regressions. All regressions include a constant, interview noise controls and country dummies (not reported). Robust standard errors, clustered by 4-digit NACE sector, in parentheses. Asterisks indicate statistical significance at the 10% (*), 5%(**) and 1%(***) level.

closure does not attain very high levels for the average firm (cf. columns 2 and 3 of Table 3). This result holds up when the regression is weighted by employment. In fact, the coefficient estimates on groups **A** and $B \cap CI > 5$ both become stronger, indicating that some of the larger firms in those categories are at a higher risk.

In order to account for the qualitative difference between a slight increase in downsizing risk and a strong downsizing impact, we also estimate Probit regressions of the binary event that a firm has a VS of 3 or larger. The results, reported in columns 4 and 5 of Table 3, confirm that only groups **A** and $B \cap CI > 5$ present some risk of downsizing. It would therefore seem justified to adjust the thresholds for exemption accordingly.

4.2.2 Refining the trade intensity definition

The evidence shows that the TI criterion is of limited value for proxying a sector's actual downsizing risk. One reason for this could be that this indicator is not precise enough to capture how exposure to international markets affects downsizing risk. For example, being exposed to competition from China might affect a firm's competitiveness in a very different way than

Table 4: Regressions of the vulnerability score on CI and region specific TI

	(1)	(2)	(3)
	Vulnerability Score		
Sectoral Carbon Intensity (CI)	0.234*** (0.060)	0.547*** (0.169)	0.551*** (0.166)
Sectoral Trade Intensity (TI)	0.376** (0.164)	0.695*** (0.232)	1.454*** (0.245)
with LESS developed countries			
TI with LEAST developed countries	-0.228*** (0.076)	-0.422*** (0.157)	-0.740*** (0.174)
TI with Developed non-EU countries	0.117 (0.125)	-0.216 (0.243)	-0.593*** (0.219)
TI with EU countries	-0.229** (0.114)	-0.411*** (0.143)	-0.680*** (0.190)
CI X CI		-0.069** (0.030)	-0.092** (0.045)
TI less X TI less		-0.154 (0.121)	-0.718*** (0.131)
TI least X TI least		0.047* (0.027)	0.094*** (0.029)
TI developed X TI developed		0.074 (0.088)	0.212*** (0.074)
TI EU X TI EU		0.014 (0.091)	0.305*** (0.110)
TI less X CI		0.378 (0.290)	0.233 (0.425)
TI least X CI		0.708*** (0.212)	0.762*** (0.187)
TI developed X CI		-0.779*** (0.232)	-0.685*** (0.179)
TI EU X CI		0.167 (0.173)	0.062 (0.223)
Weights	no	no	employment
Observations	389	389	389

Notes: OLS regressions in columns 1 and 2. WLS regression in column 3. The dataset is a cross section of 389 interviewed firms that are part of the EU ETS and for which CITL data, carbon intensity data and geographically precise sectoral trade and carbon intensity data are available. Robust standard errors, clustered by 4-digit NACE sector, in parentheses. Asterisks indicate statistical significance at the 10%(*), 5%(**) and 1%(***) level. Includes a constant, country dummies and interview noise controls (not reported). The dependent variable is the vulnerability score of the firm given by the interviews data. As explanatory variables, CI indicates carbon intensity and TI trade intensity which are calculated from Eurostat and the EU Commission data. X indicates that the two variables are interacted or squared.

does competition from Australia. Moreover, being export intensive could have different implications than being import intensive. In order to explore whether a refined TI measure would give a better indicator of carbon leakage risk, we regress VS on CI and four separate measures of the intensity of trade with (i) least developed countries (according to the UN classification), (ii) less developed (or developing) countries including China and India, (iii) developed non-EU countries and (iv) EU countries.

Table 4 summarizes the results of these regressions. Column 1 reveals a strong positive association between vulnerability and TI with less developed countries, which includes China and other countries that tend to have less stringent environmental regulation standards and which compete with European manufacturing firms. The relationship between vulnerability and TI with least developed countries is negative and significant. This could reflect a lack of competition from such countries as they tend to export agricultural products and natural resources rather than manufactured goods. High TI with EU countries is negatively associated with the VS. This is consistent with firms anticipating that their EU competitors will be subject to the same policy constraints. The findings obtained in the quadratic form, which includes interactions of TI with CI and squared terms, are qualitatively similar (column 2). In addition, TI with other developed countries outside the EU only matters in interaction with high CI, in which case vulnerability is lower. Conversely, the negative link between vulnerability and TI for the least developed countries is partially offset for the most carbon intensive firms. The employment-weighted regression shows that especially the large firms in sectors that have a high TI with less developed countries are relatively more at risk of downsizing (column 3). The coefficients on TI with other regions are negative.

In further specifications, reported in Table 5, we decompose the TI measure into export intensity (EI) and import intensity (II). This does not yield more significant results than for the overall TI measure. After differentiating trade intensities by region as above, we find that exports and imports to less developed countries are both positively associated with VS, confirming our earlier finding for overall trade with these regions.

Table 5: Regressions of the vulnerability score on CI, EI and II

	(1)	(2)	(3)
	Vulnerability Score		
Carbon Intensity (CI)	0.217*** (0.058)	0.611*** (0.201)	0.312 (0.202)
Sectoral Export intensity (EI)	-0.072 (0.160)		
Sectoral Import intensity (II)	0.142 (0.153)		
EI with LESS developed countries		0.200 (0.263)	1.613*** (0.286)
II with LESS developed countries		0.350 (0.225)	0.640** (0.273)
EI with LEAST developed countries		-0.476** (0.203)	-0.833*** (0.240)
II with LEAST developed countries		0.030 (0.185)	-0.052 (0.284)
EI with Developed non-EU countries		-0.083 (0.242)	-0.551** (0.216)
II with Developed non-EU countries		-0.156 (0.416)	-0.443 (0.374)
EI with EU countries		0.544 (0.544)	0.016 (0.675)
II with EU countries		-0.827 (0.579)	-0.901 (0.682)
EI less X EI less		0.081 (0.164)	-0.467*** (0.171)
II less X II less		-0.018 (0.102)	-0.363** (0.139)
EI least X EI least		0.089*** (0.034)	0.095 (0.097)
II least X II least		-0.012 (0.020)	0.007 (0.034)
EI developed X EI developed		0.328** (0.137)	0.303** (0.134)
II developed X II developed		-0.044 (0.098)	0.010 (0.110)
EI EU X EI EU		-0.926** (0.361)	-0.243 (0.386)
II EU X II EU		0.633** (0.305)	0.695** (0.295)
EI less X CI		-0.027 (0.386)	0.918* (0.512)
II less X CI		0.262 (0.214)	0.191 (0.295)
EI least X CI		0.255 (0.257)	0.145 (0.262)
II least X CI		0.064 (0.169)	0.411* (0.233)
EI developed X CI		0.311 (0.338)	0.153 (0.394)
II developed X CI		-0.354 (0.292)	-1.218*** (0.410)
EI EU X CI		0.041 (1.419)	-3.959** (1.659)
II EU X CI		0.158 (1.479)	3.700** (1.692)
Weights	no	no	employment
Observations	389	389	389

Notes: OLS regressions in columns 1 and 2. WLS in column 3. The dataset is a cross-section of 389 interviewed firms that are part of the EU ETS for which CITL, geographically precise sectoral trade and carbon intensity data are available. Robust standard errors, clustered by 4-digit NACE sector, in parentheses. Asterisks indicate statistical significance at the 10%(*), 5%(**) and 1%(***) level. Includes a constant, country dummies and interview noise controls (not reported). The dependent variable is the vulnerability score of the firm given by the interview data. In column 3, the firm's employment is used to weight the regression. As explanatory variables, CI indicates carbon intensity, EI export intensity and II import intensity which are calculated from Eurostat and the EU Commission data. X indicates that the two variables are interacted or squared.

5 Discussion

Our analysis of the correlation between a measure of carbon leakage based on managers' responses, and the carbon leakage criteria applied by the EC has revealed that carbon intensity is a good proxy for leakage risk whereas trade intensity is not. This mismatch gives rise to overly generous compensation in the form of free permits granted to trade-exposed industries that are not really at risk of relocating due to carbon pricing. Based on our results, we have identified two simple modifications of the carbon leakage criteria that might inform the EC's impending review of carbon leakage sectors, scheduled for 2014. The first one is to consider trade intensive sectors at risk only if they are also carbon intensive. The second modification suggests the adoption of a more specific TI measure which is based on trade only with less developed countries rather than with all non-EU countries.

If exemptions from permit auctioning were granted according to these modified criteria, more emission permits could be auctioned without a significant increase in leakage risk. Given the scale of the EU ETS, it is worthwhile to perform a back-of-the-envelope calculation of the resulting increase in auction revenue. To this end, we compile installation-level data on benchmarking allocations, available for 22 countries, and match in the information on the NACE industry code, which is needed to assign installations to exemption groups. The amount of emissions no longer exempt from auctioning under an alternative rule is computed taking into account that installations in non-exempt sectors get free permits for only 80% instead of 100% of their benchmark emissions in 2013, and that this proportion falls linearly to 30% until 2020. The results of this exercise are reported in Table 6.

Consider first a modification of the trade intensity threshold, as described above. Table 7 lists all sectors that would cease to be exempt from auctioning under this proposal. For our sample, we calculate that this would yield an additional 82.3 million emission rights to be auctioned. The bootstrapped confidence intervals at the 95% level indicate that the sampling error surrounding this estimate interval is quite small. The point estimate is a lower bound as it does not include (i) a small proportion of installations that could not be matched to industry codes and (ii) installations in seven countries for which the NIM data were not publicly available.⁹ Us-

⁹The Czech Republic, Latvia, Liechtenstein, Lithuania, Malta, Norway and Slovenia.

Table 6: Reduction of free permit allocation and additional revenue

	Reduction of free permit allocation (22 countries) [MtCO ₂ eq]	Reduction of free permit allocation, whole EU ETS [MtCO ₂ eq]	Additional revenue with price of €30 per ton [M€]	Additional revenue with price of €5 per ton [M€]
<i>A and B & CI>5</i>	82.27 [70.78; 95.54]	100.29 [86.51; 115.54]	3,008.78 [2,595.35; 3,466.17]	501.46 [432.56; 577.69]
<i>A, B and C – but TI with less developed countries only</i>	8.29 [6.32; 10.17]	14.35 [9.47; 20.11]	430.4 [284.11; 603.41]	71.73 [47.35; 100.57]

Notes: Each row reports the reduction of free permit allocations and additional revenue under a different rule. A, B and C refer to the EU criteria defined in Section 2.1. The second row uses trade intensity (TI) with less developed countries in the definition of groups B and C. MtCO₂eq stands for million metric tons of CO₂ equivalent. The numbers in brackets report two-sided 95% confidence intervals of the reductions of free permit allocation and additional revenue obtained from a bootstrap with 200 replications.

ing aggregate data on emissions in 2009, we scale up the initial estimate to the entire EU ETS and obtain a total of 100.3 million permits to be auctioned.¹⁰ Finally, we translate emissions into revenues using two alternative allowance prices. The higher price of €30 is considered in keeping with the EU Commission (2009). A lower price of €5 is closer to the market price observed during 2012 and 2013. This leads to an estimate of additional auction revenue of either €0.5 billion or €3 billion per year, with uncertainty of ±15%.

When the exemption categories are maintained but TI with less developed countries is used instead of overall TI, the increase in auction revenue is lower, €71 million to €430m million per year, depending on the allowance price, and estimated somewhat less precisely. While these revenue estimates are also subject to uncertainty about future carbon emissions and allowance prices, their order of magnitude shows that the EU is prepared to hand out profit subsidies to polluting firms on an enormous scale without getting anything in return.

6 Conclusion

In spite of a push towards more permit auctioning during the current third trading phase of the EU ETS, the evidence presented in this paper substantiates concerns that the European Commission compensates polluting industries too generously at the expense of European taxpayers.

¹⁰This is done in two steps. First, for each CITL sector in each of the 22 countries, extra auctioning is scaled up by the proportion of matched 2009 allocations for the respective sector-country pair. Second, for each CITL sector, additional auctioned permits were divided by the share of the 22 countries in the total, EU ETS wide allocation for that sector in 2009.

Table 7: List of additional sectors *not* to be exempted from auctioning

<i>Sector Description</i>	<i>NACE sector code (Rev 1.1)</i>	<i>Sector Description</i>	<i>NACE sector code (Rev 1.1)</i>
Processing and preserving of fish and fish products	152	Manufacture and processing of other glass including technical glassware	2615
Manufacture of crude oils and fats	1541	Manufacture of non-refractory ceramic goods other than for construction purposes; manufacture of refractory ceramic	262
Manufacture of starches and starch products	1562	Manufacture of ceramic tiles and flags	263
Manufacture of sugar	1583	Production of abrasive products	2681
Manufacture of distilled potable alcoholic beverages	1591	Manufacture of tubes	272
Production of ethyl alcohol from fermented materials	1592	Precious metals production	2741
Manufacture of wines	1593	Lead, zinc and tin production	2743
Manufacture of other non-distilled fermented beverages	1595	Manufacture of cutlery	2861
Preparation and spinning of woollen-type fibres	1712	Manufacture of tools	2862
Preparation and spinning of worsted-type fibres	1713	Manufacture of fasteners, screw machine products, chain and springs	2874
Preparation and spinning of flax-type fibres	1714	Manufacture of other fabricated metal products, n.e.c.	2875
Throwing and preparation of silk, including from noils, and throwing and texturing of synthetic or artificial filament yarns	1715	Manufacture of machinery for the production and use of mechanical power, except aircraft, vehicle and cycle engines	291
Manufacture of sewing threads	1716	Manufacture of furnaces and furnace burners	2921
Preparation and spinning of other textile fibres	1717	Manufacture of non-domestic cooling and ventilation equipment	2923
Textile weaving	172	Manufacture of other general purpose machinery n.e.c.	2924
Manufacture of made-up textile articles, except apparel	174	Manufacture of agricultural and forestry machinery	293
Manufacture of other textiles	175	Manufacture of machine- tools	294
Manufacture of knitted and crocheted fabrics	176	Manufacture of other special purpose machinery	295
Manufacture of knitted and crocheted articles	177	Manufacture of weapons and ammunition	296
Manufacture of other wearing apparel and accessories	182	Manufacture of electric domestic appliances	2971
Dressing and dyeing of fur; manufacture of articles of fur	183	Manufacture of office machinery and computers	300
Tanning and dressing of leather	191	Manufacture of electric motors, generators and transformers	311
Manufacture of luggage, handbags and the like, saddlery and harness	192	Manufacture of electricity distribution and control apparatus	312
Manufacture of footwear	193	Manufacture of insulated wire and cable	313
Sawmilling and planing of wood, impregnation of wood	201	Manufacture of accumulators, primary cells and primary batteries	314
Manufacture of articles of cork, straw and plaiting materials	2052	Manufacture of lighting equipment and electric lamps	315
Manufacture of pulp, paper and paperboard	211	Manufacture of other electrical equipment n.e.c.	3162
Manufacture of wallpaper	2124	Manufacture of electronic valves and tubes and other electronic components	321
Other publishing	2215	Manufacture of television and radio transmitters and apparatus for line telephony and line telegraphy	322
Manufacture of refined petroleum products	232	Manufacture of television and radio receivers, sound or video recording or reproducing apparatus and associated goods	323
Processing of nuclear fuel	233	Manufacture of medical and surgical equipment and orthopaedic appliances	331
Manufacture of dyes and pigments	2412	Manufacture of instruments and appliances for measuring, checking, testing, navigating and other purposes, except industrial process control equipment	332
Manufacture of pesticides and other agro-chemical products	242	Manufacture of optical instruments and photographic equipment	334
Manufacture of pharmaceuticals, medicinal chemicals and botanical products	244	Manufacture of watches and clocks	335
Manufacture of perfumes and toilet preparations	2452	Building and repairing of ships and boats	351
Manufacture of essential oils	2463	Manufacture of aircraft and spacecraft	353
Manufacture of photographic chemical material	2464	Manufacture of motorcycles and bicycles	354
Manufacture of prepared unrecorded media	2465	Manufacture of other transport equipment n.e.c.	355
Manufacture of other chemical products n.e.c.	2466	Manufacture of jewellery and related articles	362
Manufacture of man-made fibres	247	Manufacture of musical instruments	363
Manufacture of rubber tyres and tubes	2511	Manufacture of sports goods	364
Manufacture of flat glass	2611	Manufacture of games and toys	365
Manufacture of hollow glass	2613	Miscellaneous manufacturing n.e.c.	366

Notes: The table lists sectors that will be exempted from auctioning under the current EC criteria, but would no longer be exempted under our proposed rule change. The list contains about half of the sectors currently exempted under EU Commission proposals. The EC criteria apply at the 4 digit (NACE Rev. 1.1) sectoral level. For conciseness, we report the 3-digit sector if all 4-digit sub sectors in a 3-digit sector would cease to be exempted.

Clearly, subsidizing “carbon fat cats”¹¹ in times of deep cuts in public spending could (further) undermine political support for emissions trading. However, our analysis also points to a window of opportunity for European governments to improve the design of the EU ETS significantly while raising additional revenue in the billions of euros annually. Rather than providing an unspecific subsidy for industry, governments could earmark this money to finance investments in infrastructure and R&D which are costly but crucial for the transition to a low-carbon economy. Furthermore, part of the additional revenue could be used to mitigate possibly regressive effects of higher carbon prices on low-income groups. Not least, more permit revenue would help to balance strained government budgets in those European countries most affected by the grand recession.

References

- Abrell, J., Ndoye, A., and Zachmann, G. (2011). Assessing the impact of the EU ETS using firm level data. Bruegel Working Paper 2011/08, Brussels, Belgium.
- Anger, N., and Oberndorfer, U. (2008). Firm performance and employment in the EU emissions trading scheme: An empirical assessment for Germany. *Energy Policy*, 36(1), 12–22.
- Bloom, N., and van Reenen, J. (2007). Measuring and Explaining Management Practices across Firms and Countries. *Quarterly Journal of Economics*, CXXII(4), 1351–1406.
- Bloom, N., and van Reenen, J. (2010). New Approaches to Surveying Organizations. *American Economic Review*, 100(2), 105–09.
- Bushnell, J. B., Chong, H., and Mansur, E. T. (2013). Profiting from Regulation: An Event Study of the EU Carbon Market. *American Economic Journal: Economic Policy*, 5(4).
- Chan, H. S., Li, S., and Zhang, F. (2013). Firm competitiveness and the European Union emissions trading scheme. *Energy Policy*, (0).
- Clò, S. (2010). Grandfathering, auctioning and Carbon Leakage: Assessing the inconsistencies of the new ETS Directive. *Energy Policy*, 38(5), 2420–2430.

¹¹<http://www.sandbag.org.uk/maps/companymap/>

- Commins, N., Lyons, S., Schiffbauer, M., and Tol, N. C. (2011). Climate Policy and Corporate Behavior. *The Energy Journal*, 32(4).
- Convery, F. J., and Redmond, L. (2007). Market and Price Developments in the European Union Emissions Trading Scheme. *Review of Environmental Economics and Policy*, 1(1), 88–111.
- Demaily, D., and Quirion, P. (2006). CO2 abatement, competitiveness and leakage in the European cement industry under the EU ETS: grandfathering versus output-based allocation. *Climate Policy*, 1, 93–113.
- Demaily, D., and Quirion, P. (2008). European Emission Trading Scheme and competitiveness: A case study on the iron and steel industry. *Energy Economics*, 30(4), 2009–2027.
- Ellerman, A. D., Buchner, B. K., and Carraro, C. (Eds.) (2007). *Allocation in the European Emissions Trading Scheme: Rights, Rents and Fairness*. Cambridge: Cambridge University Press.
- Ellerman, A. D., and Joskow, P. L. (2008). The European Union’s Emissions Trading System in Perspective. Tech. rep., Pew Center on Global Climate Change, Washington, DC.
- EU Commission (2009). Impact Assessment accompanying the commission decision determining a list of sectors and subsectors which are deemed to be exposed to a significant risk of carbon leakage pursuant to article 10a (13) of Directive 2003/87/ec.
- Hahn, R. W. (1984). Market Power and Transferable Property Rights. *The Quarterly Journal of Economics*, 99(4), 753–765.
- Hood, C. (2010). Reviewing Existing and Proposed Emissions Trading Systems. Tech. Rep. 13, International Energy Agency, Paris.
- Juergens, I., Barreiro-Hurlé, J., and Vasa, A. (2013). Identifying carbon leakage sectors in the EU ETS and implications of results. *Climate Policy*, 13(1), 89–109.
- Lecourt, S., Pallière, C., and Sartor, O. J. (2013). The impact of emissions-performance benchmarking on free allocations in EU ETS Phase 3. Tech. Rep. 2013-02, Paris Dauphine CDC Climat.

- Martin, R., de Preux, L. B., and Wagner, U. J. (2011). The Impacts of the Climate Change Levy on Manufacturing: Evidence from Microdata. Working Paper 17446, National Bureau of Economic Research.
- Martin, R., Muûls, M., de Preux, L. B., and Wagner, U. J. (2012). Anatomy of a Paradox: Management Practices, Organizational Structure and Energy Efficiency. *Journal of Environmental Economics and Management*, 63(2), 208–223.
- Martin, R., Muûls, M., de Preux, L. B., and Wagner, U. J. (2013a). Industry Compensation Under Relocation Risk: A Firm-Level Analysis of the EU Emissions Trading Scheme. Working Paper 19097, National Bureau of Economic Research.
- Martin, R., Muûls, M., de Preux, L. B., and Wagner, U. J. (2014). Industry compensation under relocation risk: A firm-level analysis of the eu Emissions Trading Scheme. *American Economic Review*, forthcoming.
- Martin, R., Muûls, M., and Wagner, U. J. (2013b). The Impact of the EU ETS on Regulated Firms: What is the Evidence After Eight Years? Mimeograph.
- Meunier, G., Ponssard, J.-P., and Quirion, P. (2012). Carbon Leakage and Capacity-Based Allocations. Is the EU right? CESifo Working Paper Series 4029, CESifo Group Munich.
- Monjon, S., and Quirion, P. (2011). Addressing leakage in the EU ETS: Border adjustment or output-based allocation? *Ecological Economics*, 70(11), 1957–1971.
- Montgomery, W. (1972). Markets in licenses and efficient pollution control programs. *Journal of Economic Theory*, 5(3), 395–418.
- Paltsev, S. V. (2001). The Kyoto Protocol: Regional and Sectoral Contributions to the Carbon Leakage. *The Energy Journal*, 22(4), 53–79.
- Reguant, M., and Ellerman, A. D. (2008). Grandfathering and the endowment effect - an assessment in the context of the Spanish national allocation plan. MIT Center for Energy and Environmental Policy Research Paper 0818, Cambridge, MA.

- Reinaud, J. (2005). Industrial competitiveness under the European Union Trading Scheme. Tech. rep., International Energy Agency, Paris.
- [Sato, M., Grubb, M., Cust, J., Chan, K., Korppoo, A., and Ceppi, P. \(2007\). Differentiation and dynamics of competitiveness impacts from the EU ETS. CWPE 0712, Faculty of Economics, University of Cambridge, UK.](#)
- [Smale, R., Hartley, M., Hepburn, C. J., Ward, J., and Grubb, M. \(2006\). The impact of CO2 emissions trading on firm profits and market prices. *Climate Policy*, 6\(1\), 31–48.](#)
- Stavins, R. N. (1995). Transaction Costs and Tradeable Permits. *Journal of Environmental Economics and Management*, 29(2), 133–148.
- Wagner, U. J., Muûls, M., Martin, R., and Colmer, J. (2013a). An evaluation of the impact of the EU emissions trading system on the industrial sector. Plant-level evidence from France. Tech. rep., Imperial College London, London, UK.
- [Wagner, U. J., Rehdanz, K., and Petrick, S. \(2013b\). The impact of carbon trading on industry: Evidence from German manufacturing plants. Tech. rep., Universidad Carlos III de Madrid.](#)